

Biodegradation Capacity and Activity Enzymatic of *Bacillus subtilis* Against Low-Density Polyethylene

ABSTRACT

Aims: The objective of this work was to determine the degradation capacity of low-density polyethylene by the bacterium *Bacillus subtilis*, and analyze the production of extracellular laccase activity.

Methodology: The experiments was realized in 50 mL of culture medium, added with a fragment of known dry weight (1 cm² colorless polyethylene bag squares), and were incubated at 28°C, pH 6.5, for 6 months under static conditions, determining the growth of the bacterium by dry weight (68, 75, and 91 mg), the production of extracellular protein (271, 234, and 326.1 mg/mL), and the degradation of the substrate by dry biodegraded (8.57%, 5.88%, and 11.76%).

Results: The production of extracellular laccase enzyme was analyzed in presence of polyethylene, finding an enzymatic activity of laccase of 2.06, 1.49, and 2,03 U/mL, while in the control without substrate, no enzymatic activity was observed, which suggests that this enzyme may participate in the degradation of polyethylene. In addition, some characteristics of the extracellular enzymatic activities were analyzed, such as stability at 4°C and 28°C, optimal pH and temperature, the effect of protein and substrate concentration.

Conclusion: The extracellular protein production and dry weight of the bacterium are higher in the presence of low-density polyethylene. The laccase activity is very stable at 4°C and 28°C, the most effective pH and temperature, were 4.5 and 28°C, and present an incubation time of 5 minutes, and this data suggest that this enzymatic activities may participate in the degradation of low density polyethylene.

Keywords: *Bacillus subtilis*, polyethylene, biodegradation, extracellular laccases

1. INTRODUCTION

Plastics are organic materials that are obtained through chemical reactions using different synthetic and/or natural raw materials and are part of a group of compounds called polymers. Initially, they were manufactured using polymers and vegetable resins, such as cotton cellulose, furfural from the husk of *Avena sativa*, seed oil and casein from milk, and the first fully synthetic plastic was Bakelite (1907), to replace the use of natural products, as well as obtaining a simple, inexpensive, hard, and aesthetic product, to replace other natural products that are difficult to obtain [1]. In 2017, the world production of plastics was 348 million tons, the main producers being: Asia (50.1%, with China being the largest producer with 29.4%), Europe (18.5%), North America (Mexico, the United States and Canada, 17.7%), Africa and the Middle East (7.1%), Latin America (4%), and the Commonwealth of Independent States (former Soviet Republics, 2.6%) [2], thus currently, these products are one of the world's major concerns due to the large number of environmental problems that they cause, mainly due to their excessive consumption, which when eliminated, become very difficult to eliminate waste. For example, for bottled beverages, 500 billion tons of plastic bottles are produced per year [3], and it has been described that Mexico City, "is a large body with clogged plastic veins", since the approximately 22 million inhabitants, each day produce almost 13,000 tons of solid waste, of which 123 tons are plastic waste [2] and, due to its mismanagement, as well as the custom of discarding them in streets, gardens, sewers, etc., cause an obstruction of the drainage, floods and other problems in the city, so their use worldwide is already unsustainable, so the use of plastic in daily life, and try to reuse it [2], in addition to the fact that the use of plastic containers is generally single-use [4].

24 Plastics are widely used due to their multiple applications, polyethylene being the most widely used plastic, of which two
25 types have been reported: high-density and low-density, which are in great demand worldwide. to produce plastic bags
26 that serve as packaging for food and articles of all kinds, which leads to the excessive accumulation of these plastics in
27 the world [2]. In addition, they are used in the manufacture of containers (bottles and garbage cans) [5], packaging such
28 as bags, membranes, sheets and films [6], as well as products as varied as overalls, pipes and joints for hip replacement,
29 so it is very common to see plastic debris anywhere in the world [1, 7], since these can remain in nature between
30 hundreds and thousands years [2, 8], so that today plastic waste is a serious threat on a global scale [9]. Different
31 investigations have widely documented the great negative impact that the pollution that these products cause in the world
32 [10], for example: more than thirteen million tons of plastic end up in our oceans [11] In Mexico, one out of every five fish
33 for human consumption contains microplastics in its viscera, which affects people's health and sources of work related to
34 fishing and tourism [12]. In addition, PET nanoparticles interact with the calcium ion affecting the tissue
35 contraction/relaxation function, which could affect the functioning of the intestine of rodents [13]. Also, plastic
36 contamination has been reported in Mexican protected natural areas, which shows that this type of contamination is
37 present in the Mexican Republic beyond clandestine dumps, garbage thrown in the streets and landfills full of products
38 that supposedly they must be recycled [7]. This indicates that our consumption decisions have an impact on the cleanest,
39 most remote, and protected places on the planet, and as is evident, plastic pollution on our planet negatively affects
40 biodiversity and hinders the main strategy of conservation of ecosystem services [7].

41 On the other hand, different methods of degradation of low-density plastics have been reported, which can be physical,
42 chemical, and biological. Among the physical ones are photo-degradation and thermodegradation, and of the chemical ones,
43 oxo-degradation [14]. Also, the separation of microplastics by density has been used through the application of
44 physicochemical processes with zinc chloride in wastewater collected from the public discharges of the sewerage system
45 of the city of Riobamba (Ecuador) [15]. But biodegradation is the method that is being used more exhaustively for its
46 elimination, by means of microorganisms that degrade it by means of enzymes, although this degradation takes place
47 very slowly [10]. Therefore, the use of a wide variety of microorganisms for the degradation of this type of pollutant is being
48 widely investigated, such as: The biodegradation of plastic and polypropylene with larvae of the Coleópter *T. molitor* [5],
49 *Aspergillus flavus* fungus isolated in the presence from humus and domestic composting [16], and from an orange in a
50 state of decomposition [17], the bacteria *Bacillus cereus* and *Aeromonas hydrophila* and the fungi *Penicillium* sp., and
51 *Aspergillus* sp., isolated of sanitary landfills [18], the biodegradation of low-density polyethylene by fungi and bacterial
52 consortia isolated from municipal garbage dumps [6], the biodegradation of polystyrene, PET and polyphenyl sulfide
53 plastic beads by *Pseudomonas* sp., *P. aeruginosa* and *Tichoderma* spp., [19, 20, and 21], the biodegradation capacity of
54 five filamentous fungi against polyethylene [10], the biodegradation of low-density polyethylene by a microbial consortium
55 [14], the degradation of high-density polyethylene of marine debris by *Aspergillus tubingensis* and *A. flavus* [22], the
56 biodegradation of low-density polyethylene by *Microbulbifer hydrolyticus* IRE-31 [23], the biodegradation of polyvinyl
57 chloride plastic films by a marine consortium [24] as well as the degradation of plastic by environmental bacteria in
58 Norway [25].

59 In addition, some enzymes that apparently participate in the degradation of polyethylene have been studied, which
60 hydrolyze the ester bonds, causing the release of terminal groups of carboxylic and alcoholic acids [26], like the activity
61 of laccases and esterases produced by *F. culmorum* grown in the presence of different concentrations of di (2-ethyl hexyl)
62 phthalate and Tween 80 [27], a laccase of *Trichoderma viride* [28], a recombinant laccase from *Streptomyces cyaneus*
63 CECT 3335 [29], a purified laccase from *Geobacillus* sp. ID17 [30], the esterase activity of *Pseudomonas* sp., which
64 degrades polyurethane and low-density polyethylene [31], the activity of fungal esterases on the degradation of polyesters
65 [32], an esterase from *Sphingobium* sp., C3 that degrades dimethyl terephthalate [33], two enzymatic activities of esterase
66 and phthalate hydrolase from *Gordonia* sp., which degrade phthalate esters [34], cutinases from *F. solani* and *Pichia*
67 *pastoris* [35], polyurethanases from *Pseudomonas* [36], hydrolases, lipases, and cutinases from different microorganisms
68 that degrade plastic [37], carboxylesterases [38], cutinase from *Escherichia coli* [39], PETase and MHETase from
69 *Ideonella sakaiensis* 201-F6 [40] and lipase, carboxymethylcellulose, xylanase and protease from *Alcaligenes faecalis* [41].
70 Therefore, the objective of this work was to evaluate the degradation capacity of low-density polyethylene from
71 commercial bags by the bacterium *Bacillus subtilis*, as well as to analyze some laccase enzymatic properties.

72 73 **2. MATERIAL AND METHODS**

74 75 **2.1 Strain Used**

76 The strain of *B. subtilis* was obtained from the Microbiology Laboratory of the Faculty of Chemical Sciences of the UASLP,
78 San Luís Potosí, S.L.P., México.

79 80 **2.2 Culture medium for the Degradation of Low-Density Polyethylene**

81 This medium contains (g/L): Glucose (10), yeast extract (5), KH_2PO_4 (0.6), $\text{MgSO}_4 \cdot 7\text{H}_2\text{O}$ (0.5), K_2HPO_4 (0.4), $\text{CuSO}_4 \cdot 5\text{H}_2\text{O}$
82 (0.25), $\text{FeSO}_4 \cdot 7\text{H}_2\text{O}$ (0.05), MnSO_4 (0.05), $\text{ZnSO}_4 \cdot 7\text{H}_2\text{O}$ (0.001), and 400 μL de Tween 80 [42]. Subsequently, 50 mL were
83

84 added to 125 mL Erlenmeyer flasks, as well as a disinfected plastic fragment of known dry weight (1 cm² polyethylene
85 bag squares) and sterilized by humid heat at 15 pounds (121°C) for 20 minutes. Subsequently, they were cooled to room
86 temperature, seeding 1 x10⁶ cells/mL in triplicate, and incubating for 6 months at room temperature, pH 6.5, under static
87 conditions, monitoring their growth visually every week, and adding new culture medium under sterile conditions every 3
88 weeks.

90 **2.3 Bacterium Growth by Dry Weight**

91
92 After 6 months of incubation under static conditions, the bacterial culture supernatant was harvested in a graduated tube,
93 previously weighed, and centrifuged at 3000rpm/10 min, discarding the supernatant. The cell pack was dried at 80°C, for
94 24 h, and the tube was weighed, determining the dry weight of the sample by difference, comparing the growth with a
95 control grown under the same conditions without the addition of the low-grade polyethylene fragment. All experiments
96 were performed at least 3 times in duplicate.

98 **2.4 Biodegraded Weight of Low-Density Polyethylene**

99
100 After the incubation period, the low-density polyethylene samples were taken with surgical forceps, and placed in
101 previously tared Petri dishes, washed with 2% (v/v) sodium dodecyl sulfate for 24 hours, subsequently with ethanol (70%)
102 and tridesionized water, and dried at 60°C for 24 hours, weighed and by weight difference, the biodegraded weight, and
103 the percentage of biodegradation of the sample were determined.

104
105 1) Biodegradability of the final weight of the low-density polyethylene sample was determined in milligrams, at 6 months of
106 incubation at 28°C, pH 6.5 under static conditions by the action of the bacterium *B. subtilis* using the following formula:

107 Biodegraded weight of the sample = initial weight-final weight

108 2) After obtaining the biodegraded weight of the difference from the initial weight minus the final weight, it was converted
109 to a percentage, using the following formula:

$$110 \text{ Weight loss (\%)} = \frac{\text{initial weight-final weight}}{\text{starting weight}} \times 100$$

113 **2.5 Determination of Protein**

114
115 This was determined by the method of Lowry *et al.* (1951) [43].

117 **2.6 Determination of Enzymatic Activity**

118
119 The enzymatic activity was determined spectrophotometrically in the culture supernatant, obtained from the filtration of the
120 samples.

122 **2.7 Laccase**

123
124 The reaction mixture contained 900 µL of 2 mM 2,6-dimethoxyphenol as substrate (Sigma Chemical Co.), in 0.1 M acetate
125 buffer pH 4.5, and 100 µL of enzyme extract (supernatant), incubating at 40°C for 1 minute [44], and determining the
126 laccase activity as the change in absorbance at a wavelength of 568 nm in a UV-Visible light spectrophotometer
127 (Shimadzu model 160-A), using as a reference a blank prepared with tridesionized water according to the previous
128 procedure. One unit of laccase activity was defined as the amount of enzyme that produces an increase of one
129 absorbance unit per minute in the reaction mixture [45]. Results are expressed as the average of 3 independent
130 determinations.

132 **3. RESULTS AND DISCUSSION**

134 **3.1 Bacterial Growth by Dry Weight**

135
136 The growth of the bacterium was analyzed in the presence of low-density polyethylene as a substrate, determining the dry
137 weight and the production of extracellular protein. In Figure 1, it is observed that the microorganism had a higher growth in
138 dry weight of 68, 75, and 91 mg, like control (75 mg) (which has no substrate), at 6 months of incubation, pH 6.5 at 28°C,
139 under static conditions, which indicates that polyethylene stimulates little the growth of the bacterium. The data found in
140 this work coincide with some reports in the literature, in which the growth of different microorganisms is reported in the
141 presence of different plastic substrates, such as the growth of five filamentous fungi in the presence of polyethylene [10],

greater growth with respect to the control of *Pseudomonas* sp., [19], the fungi *Mucor* sp., and *Aspergillus* sp., which increase their growth by 8.75% and 21.73% in presence of low-density polyethylene at 3 months of incubation [46], for the white rot fungus *P. ostreatus*, a growth of 619 mg was observed with 15 mg/L of tire dust, which were obtained from an industrial waste landfill located in Cartagena, Colombia [47]. Also, *A. alternata*, isolated from urban waste containers in 5 cities of the V region of Chile, demonstrated the ability to grow in different types of plastic, especially in polyurethane, polyvinyl chloride, and ethylene polyereftherate [48].

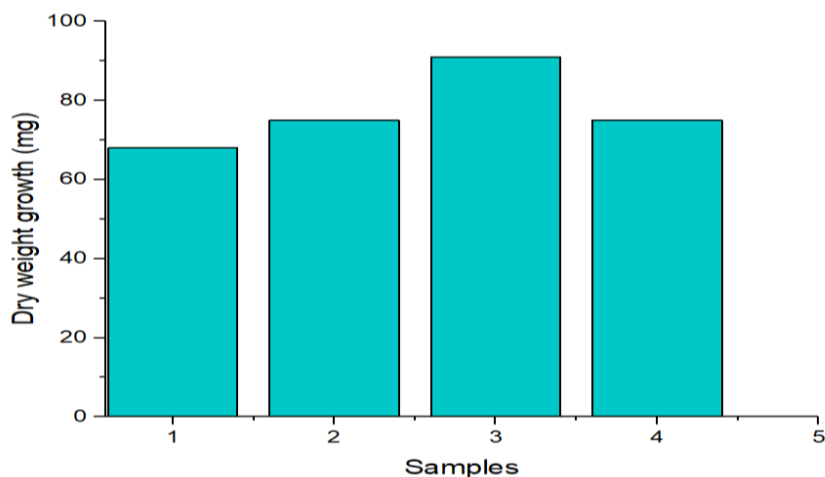


Fig. 1. Dry weight growth of *Bacillus subtilis* in presence of low-density polyethylene. 28°C. pH 6.5. 6 months of incubation. Static conditions (1×10^6 cells/mL). (1, 2, 3, problems, and 4.- control).

3.2 Extracellular Protein Production

Regarding the production of extracellular protein, a growth related to its production was found of 2.0, 1.72, and 2.4 times more than the control without substrate (Figure 2), which coincides with that reported for the fungus *F. culmorum* that produces a large amount of extracellular protein in the presence of 20 g/L of cutin [49].

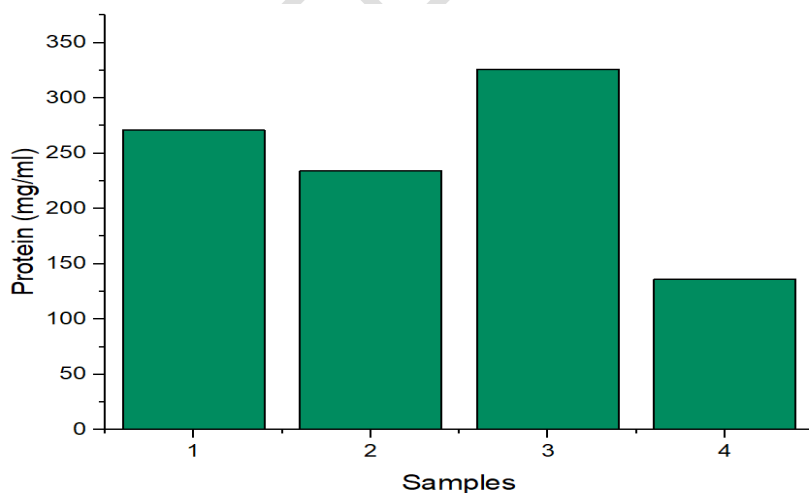


Fig. 2. Production of extracellular protein by *Bacillus subtilis* in presence of low-density polyethylene. 28°C. pH 6.5. 6 months of incubation. Static conditions (1×10^6 cells/mL). (1, 2, 3, problems, and 4.- control).

3.3 Biodegraded Weight of the Sample

In Figure 3, the biodegradation of low-density polyethylene is observed, with 8.57%, 5.88% and 11.76% of biodegradation based on the biodegraded weight of the substrate, under the conditions described above, results that coincide with that reported for three strains of the fungus *A. niger* isolated from plastic from the waste dump, from an orange in a state of decomposition, in presence of humus and domestic composting [16], which reduce 3.44%, 6.9% and 4.84% of the initial weight of polyethylene in a month, 10 days and a month, respectively [16, 17], for the fungi *Fusarium* sp., *Aspergillus* sp., *Trichoderma* sp., and *Mucor* sp., which reduce the dry weight of polyethylene from 1.0354 to 0.9533, from 1.0244 to

0.9715, from 1.096 to 0.9873, from 1.0047 to 0.9805 grams of dry weight, respectively [46]. But these results are slightly higher than the reported for the 2.88% biodegradation of low-density polyethylene by fungi and bacterial consortia isolated from municipal garbage dumps, at 70 days [6], for the 1.61% biodegradation of polystyrene at 15 days by *Pseudomonas* sp., [19] and the bacteria *P. microspora* E2712A and E3317B, which efficiently biodegrade polyurethane in liquid cultures at 16 days of incubation [50]. Also, the data found are lower than that reported for the biodegradation of the same substrate by the larvae of the Coleopter *T. molitor*, which biodegrade 64% in 45 days of incubation [5], for the bacterium *Bacillus cereus* and the fungus *Penicillium* sp., with a biodegradation of 17.91% of polyethylene terephthalate, at 4 months, although it was previously treated with UV light and thermodegradation [18], for *P. aeruginosa*, which biodegrades 21.7% and 27.3% of low-density polyethylene particles at 25°C and 35°C, respectively, after 30 days of incubation [51], and for the biodegradation of polyethylene terephthalate treated at 150°C for 8 hours, for *P. aeruginosa* (14.4%) and *Trichoderma* sp., (13.15%) during a period of 30-90 days [20].

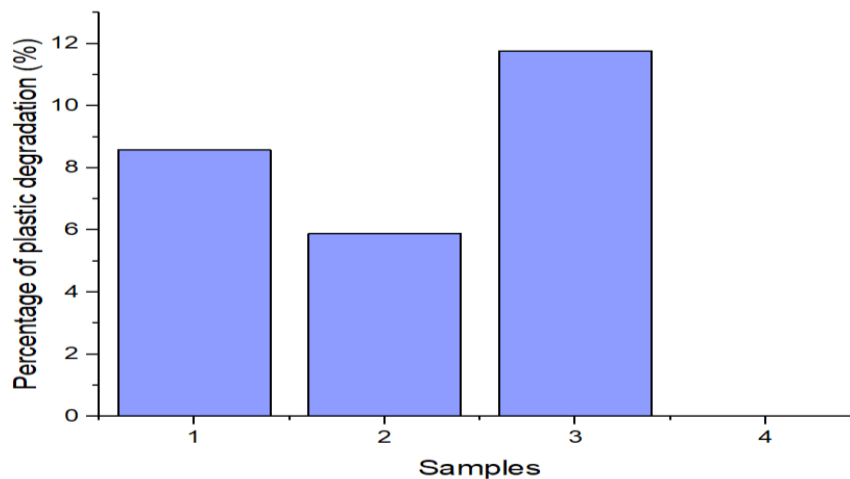
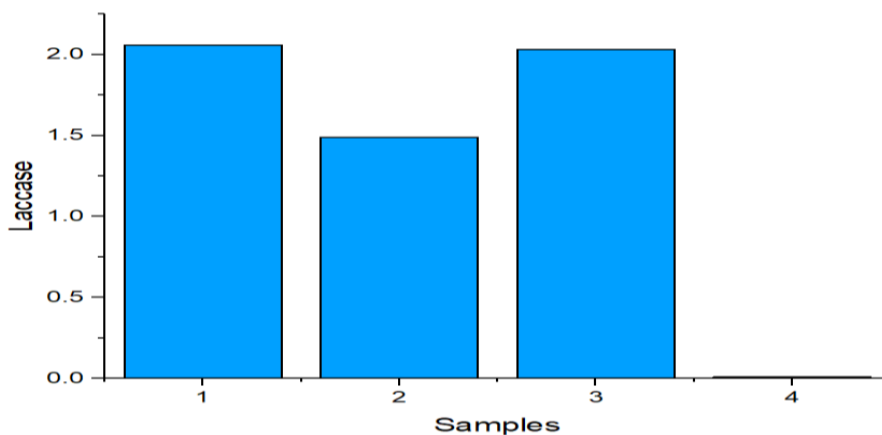


Fig.3. Percentage of biodegradation of low-density polyethylene by *Bacillus subtilis*. 28°C. pH 6.5. 6 months of incubation. Static conditions (1×10^6 cells/mL). (1, 2, 3, problems, and 4.- control).

3.4 Production of Extracellular Laccase

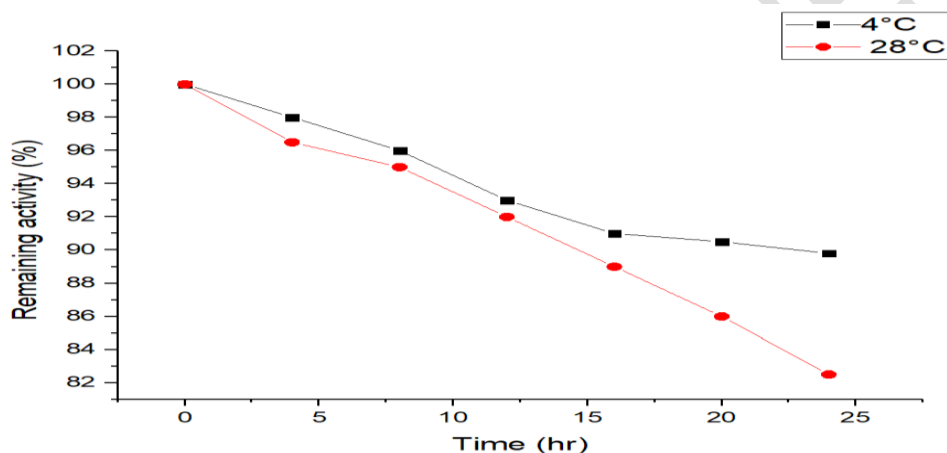
In Figure 4, the extracellular enzymatic activity of laccase produced by the bacterium *B. subtilis* is observed in presence of low-density polyurethane, under the conditions described above, finding an activity for laccase of 2.06, 1.49, and 2.03 U/mL. It should be mentioned that the controls without the substrate produced very little enzymatic activity. This is different for a laccase of *T. viride*, in which an activity of 7.31 U/mL with low-density polyurethane as substrate is reported [52], for 2 strains of *Alicyclophilus* sp., in which enzymatic activity of esterase is detected, but not of urease and protease [53], although they are lower than those reported for the production of esterase (12 U/mL) in presence of polyurethane by the bacteria *Bacillus* sp. AF8, *Pseudomonas* sp. AF9, *Micrococcus* sp. [10], *Arthrobacter* sp. AF11 and *Corynebacterium* sp. AF12 [54], a similar enzymatic activity of *F. culmorum*, where a value of 420.2 U/L is reported in the presence of 2 g/L of di (2-ethyl hexyl) phthalate at 200 hours incubation [55]. Also, for the esterase activity of different fungi isolated from sand contaminated with plastics, in which a higher esterase activity is reported with di (2-ethyl hexyl) phthalate and polyurethane foam as substrate [42].



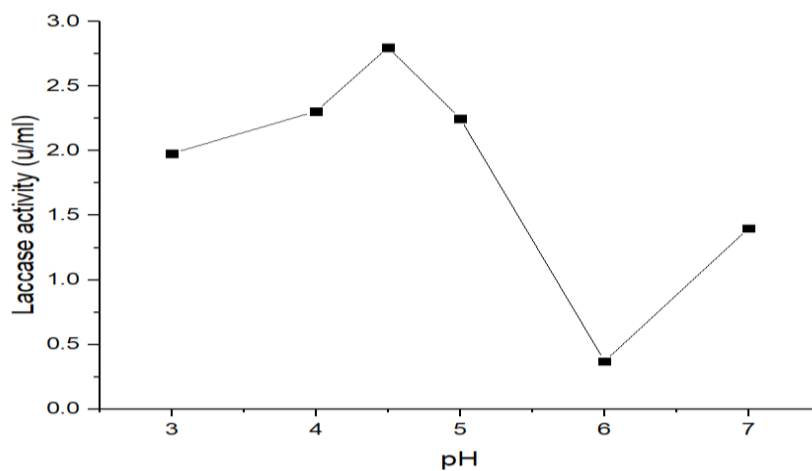
203 Fig.4.Production of extracellular laccase (U/mL) by *Bacillus subtilis* with low-density polyethylene. 28°C. pH 6.5. 6
204 months of incubation. Static conditions (1×10^6 cells/mL) (1, 2 3, Problems.4. Control.
205

206 3.5 Analysis from Some Properties of Extracellular Laccase

207
208 Subsequently, some properties of the extracellular laccase activity were analyzed. For stability, it was found that laccase
209 activity is very stable at 4°C and 28°C, conserving 90% and 82.5% of remaining activity (Figure 5), the most effective pH
210 and temperature were 4.5 (Figure 6) and 28°C (Figure 7), and an incubation time of 5 minutes (Figure 8). For the effect of
211 protein concentration, a linear reaction of laccase activity until 108.4 µg/assay of the concentrations analyzed (Figure 9),
212 while the substrate concentration (2,6-dimethoxyphenol), the highest enzyme activity was observed at 0.542 µg/assay
213 (Figure 10). In this regard, for a recombinant laccase from *S. cyaneus* CECT 3335, it has been reported that at
214 temperatures of 60°C to 80°C and pH of 3.0 the activity was greater than 75% of the maximum detected, and at
215 concentrations greater than 0.1 mM of 2,6-dimethoxyphenol, this inhibit the enzymatic activity with 2,6-dimethoxyphenol
216 substrate [29], and a purified laccase from *Geobacillus* sp. ID17, showed a similar stability at 55°C, and an optimum pH of
217 7.5[30], for a laccase from *T. viridae*, in which an optimal pH of 4.0-5.0 with low-density polyurethane as substrate, and
218 optimum temperature of 30°C and 40°C is reported [50], a carboxylesterase from *E. coli* retains 100% of its activity after 23
219 days at 45°C, and a pH of 9.0 [38], and for an extracellular depolymerase from *Penicillium oxalicum*, with an optimal
220 temperature of 40°C with aliphatic polyesters as substrates [56].
221

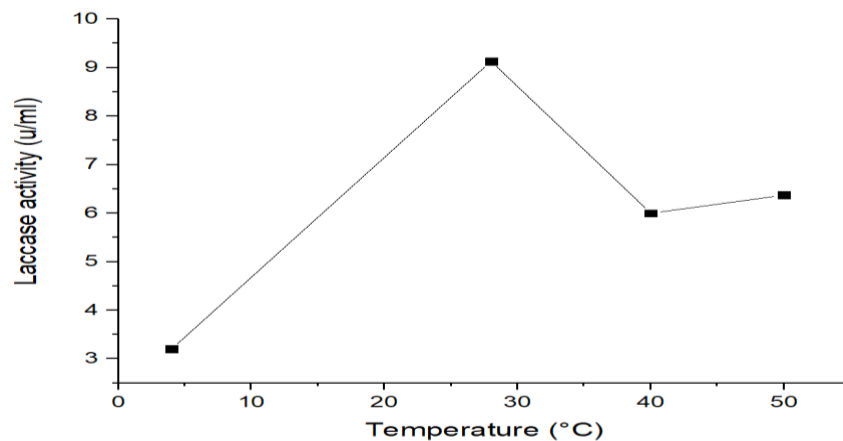


222
223 Fig.5.Stability of the laccase extracellular activity of *Bacillus subtilis* at 4°C and 28°C.
224
225
226



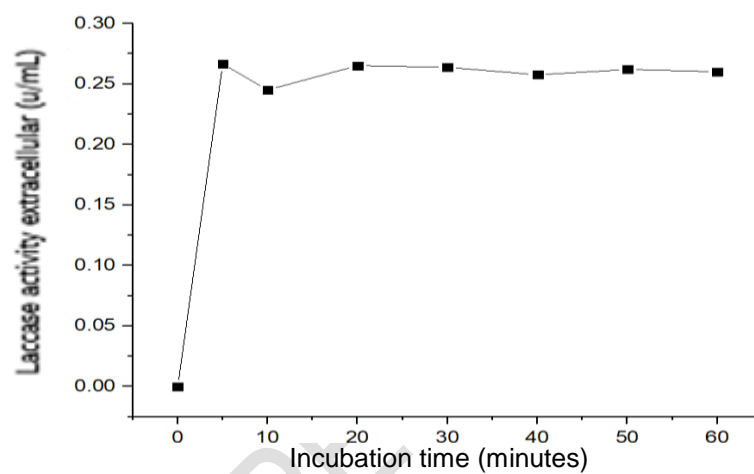
227
228 Fig.6.Effect of the pH on the laccase extracellular activity of *Bacillus subtilis* at 28°C.
229

230
231



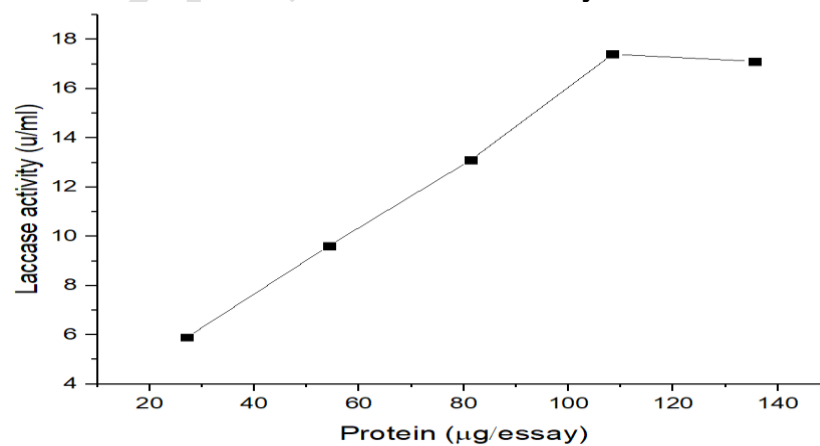
232
233
234

Fig.7.Effect of the temperature on the laccase extracellular activity of *Bacillus subtilis*.



235
236
237
238
239

Fig.8.Effect of the incubation time on the laccase extracellular activity of *Bacillus subtilis*.



240
241
242

Fig.9.Effect of the protein concentration on the laccase extracellular activity of *Bacillus subtilis*.

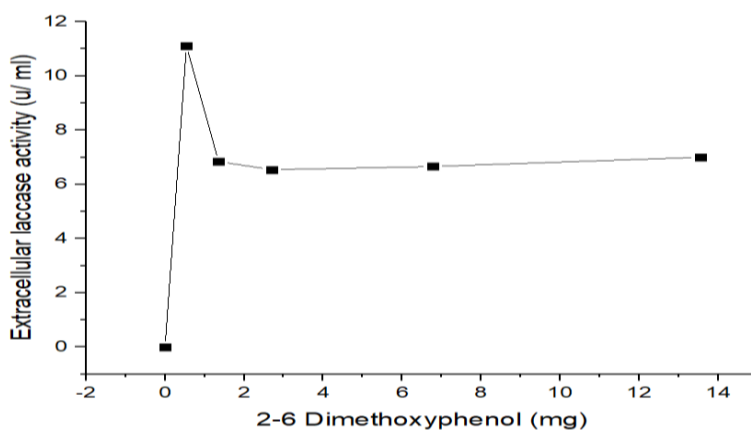


Fig.10.Effect of the 2-6 Dimethoxyphenol on the laccase extracellular activity of *Bacillus subtilis*.

Finally, a summary of the results obtained for enzyme activity is shown in Table 1.

Table 1. Kinetics characteristics of the enzymatic extracellular activity of *Bacillus subtilis*

Parameter	Laccase
Stability to 4°C	90%
Stability to 28°C	82.5
pH	4.5
Temperature	28°C
Incubation time	5 minutes
Protein concentration	108.4 µg/ensayo
Substratum concentration	0.542 µg/ensayo*

*2,6-Dimethoxyphenol

Other enzymatic activities related to the degradation of polyurethane have also been reported, such as: polyurethanases from *Pseudomonas*[36], an phthalate hydrolase from *Gordonia* sp., which degrades phthalate esters [34], hydrolases, lipases, and cutinases of different microorganisms that degrade plastic [37], carboxylesterases [38], cutinase of *E. coli*[39], PETase and MHETase from *I. sakaiensis*[40], a lipase, carboxymethylcellulose, xylanase and protease from *A. faecalis*[41], but more studies are required to determine which activities are the most efficient in the degradation of this substrate, as well as to optimize the production of the same for a faster and more efficient biodegradation

4. CONCLUSION

- 1.- The extracellular protein production and dry weight of the bacterium are higher in the presence of low-density polyethylene.
- 2.- The biodegradation of the substrate based on the biodegraded dry weight was 8.57%, 5.88%, and 11.76%.
- 3.- The bacterium produced extracellular laccase activity in presence of polyethylene, with an activity of laccase of 2.06, 1.46, and 2.03 U/mL.
- 4.- The laccase activity is very stable at 4°C and 28°C, the most effective pH and temperature, were 4.5 and 28°C, and present an incubation time of 5 minutes.
- 4.- The data obtained suggest that these enzymatic activities may participate in the degradation of low density polyethylene.

COMPETING INTERESTS

Authors have declared that no competing interests exist.

COMPETING INTERESTS DISCLAIMER:

Authors have declared that no competing interests exist. The products used for this research are commonly and predominantly use products in our area of research and country. There is absolutely no conflict of interest between the authors and producers of the products because we do not intend to use these products as an avenue for any litigation but for the advancement of knowledge. Also, the research was not funded by the producing company rather it was funded by personal efforts of the authors.

REFERENCES

- [1] Cornejo Reyes GV, MarineroOrantes EA, FunesGuadron CR, Toruño PJ, Zúñiga González CA. Biopolímeros para uso agroindustrial: Alternativas sostenibles para la elaboración de una película de algodón termoplástico biodegradable. *Rev. Iberoam. Biecon. y Cam. Clim.* 2020;6(11):1-29. <https://doi.org/10.5377/ribcc.v6i11.9824>. Español.
- [2] Martínez Arroyo MA, Ruiz Suarez LG, Gavilán García A, Mendoza Cantu A, Ramírez Muñoz T. General Panorama of the Technologies of the Recycling of plastics in Mexico and in the World. *Inst. Nac. Ecol. and Cam. Clim. SEMARNAT.* 2020. <https://www.gob.mx/uploads/attachment>. Spanish.
- [3] Lavayen Villamar KJ. Microplastic and marine pollution. Salesian Polytechnic University. Bachelor's thesis. Degree in Social Communication with a Mention in Audiovisual and Multimedia Production. 2012. Guayaquil Headquarters. Ecuador. <http://dspace.ups.edu.ec/handle/123456789/20095>. Spanish
- [4] Sánchez Durán JF. Impact of single-use plastic and alternatives for its replacement in the municipality of Urrao. Faculty of Engineering. Tecnológico de Antioquia. University institution. Bachelor's thesis. Environmental engineering. 2020. Medellín, Colombia. <https://dspace.tdea.edu>. Spanish
- [5] Álvarez Estepa DN, Botache Laguna LM. Biodegradation of plastic with larvae of the Coleoptera *Tenebrio molitor* as an interdisciplinary contribution to Environmental Biotechnology. Faculty of Science and Technology. National Pedagogical University. Thesis degree. Graduate in Biology. 2020. Bogotá, D.C. Colombia. <http://hdl.handle.net/20.500.12209/12205> Spanish
- [6] Gutiérrez Álvarez IA. Biodegradation of low-density polyethylene using fungi, bacteria and bacteria consortia isolated from the municipal dump of Tacna. Professional School of Environmental Engineering. Faculty of Engineering. Private University of Tacna. Bachelor's thesis. Environmental engineering. 2019. Tacna, Peru. <http://repositorio.upt.edu.pe/handle/UPT/1269>. Spanish
- [7] Rivera-Garibay O, Álvarez-Filip L, Rivas M, Garelli-Ríos O, Pérez-Cervantes E, Estrada-Saldívar N. Impact of plastic pollution on Mexican protected natural areas. 1st. Ed. Greenpeace Mexico. 2020. <https://www.greenpeace.org>. Spanish
- [8] Andrady AL. The plastic in microplastics: a review. *Mar. Poll. Bull.* 2017; 119 (1): 12-22. doi: 10.1016 / j.marpolbul.2017.01.082.
- [9] Barboza LGA, Cózar A, Giménez BC, Barros TL, Kershaw PJ, Guilhermino L. Macroplastics pollution in the marine environment. In Charles R. C. Sheppard (ed.), *World Seas: an environmental evaluation*, United States, Academic Press. 2019. 305-328. <http://www.oceancare.org>
- [10] González Alcos VC. Biodegradable capacity of filamentous fungi against polyethylene. *Rev. de Invest. Esc. Of Posg. Univ. Nac. Del Altip.* 2020; 9 (3): 1792-1804. DOI: <https://doi.org/10.26788/epg.v9i3.1625>. Spanish
- [11] Geyer R, Jenna R, Lavender Law J, Lavender Law K. Production, use, and fate of all plastics ever made. *Science Adv.* 2017; 3 (7): 5. DOI: 10.1126 / sciadv.1700782
- [12] Greenpeace Mexico. Study on contamination by microplastics in fish from Mexico. 2019. <https://www.greenpeace.org/mexico/publicacion/3377>. Spanish
- [13] Venegas Guerrero G. Toxicity of polyethylene terephthalate (PET) nanoparticles in a live rodent digestive system. Center for Scientific Research and Higher Education of Ensenada, Baja California. Thesis Master of Science. Nanotechnology. 2021. Ensenada, Baja California, Mexico. <http://cicese.repositorioinstitucional.mx/jspui/handle/1007/3558>. Spanish
- [14] De La Cruz Orihuela C, Arone Valencia A. Biodegradation of low-density polyethylene through a micro-microbial consortium under anaerobic and aerobic conditions. Faculty of Engineering and Architecture. Peruvian Union University. Bachelor's thesis. Environmental engineering. 2020. Lima, Peru. <http://repositorio.upeu.edu.pe/handle/UPEU/3219>. Spanish.

335 [15] Barros Barreno WA. Microplastics separation through physicochemical processes in wastewater in the city of Riobamba.
336 Faculty of Engineering. National University of Chimborazo. Bachelor's thesis. Civil Engineering. 2021. Riobamba, Ecuador.
337 <http://dspace.unach.edu.ec/handle/51000/7499>. Spanish

338 [16] Calcetero Moreno LA, Mancera Hernández JC. Evaluation of the colonization and degradation process of low-density
339 polyethylene by inoculum of *Aspergillus niger* humus and domestic composting. Faculty of Engineering. University of
340 America Foundation. Bogotá, D.C. Bachelor's thesis. Chemical engineer. 2021. Bogotá, Colombia. <https://hdl.handle.net/20.500.11839/8303>. Spanish

341 [17] Torres Herrera AA. Effectiveness of the *Aspergillus niger* fungus on the biodegradation of low-density polyethylene.
342 Faculty of Engineering and Architecture. Cesar Vallejo University. Bachelor's thesis. Environmental engineering. 2020.
343 Chiclayo, Peru.
344 <https://hdl.handle.net/20.500.12692/50246>. Spanish

345 [18] Castro Velasco AM, Avendaño Toledo CA. Determination of the most effective treatment on polyethylene terephthalate
346 to increase the efficiency of the degradation process carried out by fungi and indigenous bacteria from landfill leachate.
347 Faculty of Environmental Engineering. Socorro Free Sectional University. Bachelor's thesis. Environmental engineering.
348 2020. Santander. Colombia. <http://hdl.handle.net/10901/18618>. Spanish

349 [19] Condori Álvarez KC. Biodegradation of expanded polystyrene by *Pseudomonas* sp. isolated from the solid waste dump of
350 the city of Azángaro. Professional School of Environmental and Forest Engineering. National University of Juliaca.
351 Bachelor's thesis. Environmental and Forestry Engineering. 2020. Juliaca, Peru. <http://repositorio.unaj.edu.pe/handle/UNAJ/131>. Spanish

352 [20] Bermúdez Morera DC. Evaluation of microorganisms (*Trichoderma* sp. and *Pseudomonas aeruginosa*) for the
353 degradation of PET. Faculty of Engineering. University of America Foundation. Bogotá, D.C. Bachelor's thesis. Chemical
354 engineer. 2021. Bogotá, Colombia. <http://repositorio.uamerica.edu.co>. Spanish

355 [21] Li S, Wei R, Gao M, Ren Y, Yu B, Nie K, Xu H, Liu L.
356 Biodegradation of low density polyethylene by *Microbulbifer hydrolyticus* IRE-31. *J. Environ. Manag.* 2020; 263: 1-13. doi:
357 10.1016/j.jenvman.2020.110402

358 [22] Devi RS, Kannan VR, Nivas D, Kannan K, Chandru S, Antony AR. Biodegradation of HDPE by *Aspergillus* spp. from
359 marine ecosystem of Gulf of Mannar, India. *Mar. Poll. Bull.* 2020; 96 (1,2): 32-40. doi: 10.1016/j.marpolbul.2015.05.050.

360 [23] Li J, Kim HR, Lee HM, Yu Ch, Jeon E, Lee S, Kim D. Rapid
361 biodegradation of polyphenylene sulfide plastic beads by *Pseudomonas* sp. *Scien. Tot. Environ.* 2020; 729: 1-11, 137616. doi:
362 10.1016/j.scitotenv.2020.137616.

363 [24] Giacomucci L, Raddadi GN, Soccio M, Lotti N, Fava F.
364 Biodegradation of polyvinyl chloride plastic films by enriched anaerobic marine consortia. *Mar. Environ. Res.* 2020; 158,
365 104949. doi: 10.1016/j.marenvres.2020.104949.

366 [25] Charnock C. Norwegian soils and waters contain mesophilic, plastic-degrading bacteria. *Microorganisms.*
367 2021; 9(94): 1-18. doi: 10.3390/microorganisms9010094.

368 [26] Liu J, He J, Xue R, Xu B, Qian X, Xin F et al., Biodegradation and up-cycling of polyurethanes: Progress, challenges,
369 and prospects. *Biotechnol. Adv.* 2021; 48: 1-12. 107730. doi: 10.1016/j.biotechadv.2021.107730.

370 [27] Medina-Flores H, González-Márquez A, Sánchez C. Effect of surfactant Tween 80 on growth and esterase production
371 of *Fusarium culmorum* in liquid fermentation. *Mex. J. Biotech.* 2020; 5(4): 64-79. <https://doi.org/10.29267/mxjb.2020.5.4.64>

372 [28] Johnnie DA, Isaac R, Prabha ML. Bio efficacy assay of laccase isolated and characterized from *Trichoderma viride*
373 in biodegradation of low density polyethylene (LDPE) and textile industrial effluents dyes. *J. of Pure Appl. Microbiol.*
374 2021; 15(1): 410-420. DOI: 10.22207/JPAM.15.1.38

375 [29] Moya Lobo R. Caracterización de la lacasa de *Streptomyces cyaneus* CECT 3335 y aproximación al estudio de
376 su potencial oxidativo y función biológica. Universidad de Alcalá. Tesis Doctoral. Depto. de Microbiología y Parasitología.
377 2021. Alcalá de Henares, España. Español

378 [30] Atalah Zuñiga JI. Purificación y caracterización de una nueva lacasa aislada del microorganismo termófilo *Geobacillus*
379 sp. ID17. Facultad de Ciencias Químicas y Farmacéuticas. Universidad de Chile. Título de Bioquímico. 2017. Santiago,
380 Chile. <http://repositorio.uchile.cl/handle/2250/149820>. Español

381 [31] Roy R, Mukherjee G, Das Gupta A, Tribedi P, Kamal A. Isolation of a soil bacterium for remediation of polyurethane
382 and low-density polyethylene: A promising tool toward sustainable cleanup of the environment. *3 Biotech.* 2021; 11(29): 1-
383 13. doi: 10.1007/s13205-020-02592-9.

384 [32] Weinberger S, Beyer R, Schüller Ch, Strauss J, Pellis A, Ribitsch AD, Guebitz GM. High throughput screening for new
385 fungal polyester hydrolyzing enzymes. *Front. in Microbiol.* 2020; 11(558): 1-8. <https://doi.org/10.3389/fmicb.2020.00554>

386 [33] Cheng X, Dong SS, Chena D, Rui Q, Guo J, Wang D, Jiang J. Potential of esterase DmtH in transforming plastic additive
387 dimethyl terephthalate to less toxic monomethyl terephthalate. *Ecotox. Environ. Saf.* 2020; 187(5): 109848. doi:
388 10.1016/j.ecoenv.2019.109848

389 [34] Huang H, Zhang XY, Chen TL, Zhao YL, Xu DS, Bai YP. Biodegradation of Structurally Diverse Phthalate Esters by a
390 Newly Identified Esterase with Catalytic Activity toward Di(2-ethylhexyl) Phthalate". *J. Agric. Food Chem.* 2019; 67(31): 8548-
391 8558. doi: 10.1021/acs.jafc.9b02655

394 [35] Peña-Montes C, Bermúdez-García E, Morales-García S, Farrés A. Las cutinas como una herramienta valiosa para la
395 descontaminación de residuos plásticos. *Mensaje Bioquímico*. Vol. 42, pp. 24-35. 2018. <http://tab.facmed.unam.mx>. Español
396 [36] Petri do Canto V, Thompson CE, Netzi PA. Computational studies of polyurethanases from *Pseudomonas*. *J. Mol. Mod.*
397 2021;27(46):1-8. DOI: 10.1007/s00894-021-04671-x
398 [37] Soriano Ortega B. Biodegradación de plásticos en ambientes naturales. Facultad de Ciencias. Universidad de Alcalá.
399 Alcalá de Henares. Trabajo de Fin de Grado. Ciencias Ambientales. 2020. Madrid, España.
400 <http://hdl.handle.net/10017/45807>. Español
401 [38] Ding J, Zhou Y, Wang Ch, Peng Z, Mu Y, Tang X, Huang Z. Development a whole-cell biocatalytic for
402 diisobutylphthalate degradation by functional display of carbobylesterase on the surface of *Escherichia coli*. *Microb. Cell Fact.*
403 2020;19(114):1-11. DOI: <https://doi.org/10.1186/s12934-020-01373-6>
404 [39] Falkestein P, Gräsing D, Byelytskyi P, Zimmermann W, Matysic J, Wei R, Song Ch. Uv treatment impairs the enzymatic
405 degradation of polyethylene terephthalate. *Front. in Microb.* 2020;11(689):1-10. <https://doi.org/10.3389/fmicb.2020.00689>
406 [40] Maity W, Maity S, Brera S, Roy A. Emerging roles of PETase and MHETase in the biodegradation of plastic wastes.
407 *Appl. Biochem. Biotech.* 2021;193(8):2699-2716. doi: 10.1007/s12010-021-03562-4.
408 [41] Nag M, Lahiri D, Dutta D, Jadav G, Ray RR. Biodegradation of use polyethylene bags by a new marine strain of
409 *Alcaligenes faecalis* LND-1. *Environ. Scien. Poll. Res.* 2021;28(30):41363-41379. doi: 10.1007/s11356-021-13704-0.
410 [42] Ahuactzin-Pérez M, Tecuítl-Beristain S, García-Dávila J, González-Pérez M, Gutiérrez-Ruiz MC, Sánchez C.
411 Degradation of di(2-ethyl hexyl) phthalate by *Fusarium culmorum*: Kinetics, enzymatic activities and biodegradation
412 pathway based on quantum chemical modeling pathway. *Scien. Total Environ.* 2016;566-567:1186-1193. doi:
413 10.1016/j.scitotenv.2016.05.169.
414 [43] Lowry OH, Rosebrough NJ, Farr AL, Randall RJ. Protein measurement with the Folin phenol reagent. *J. Biol. Chem.*
415 1951;193(19):265-275. PMID: 14907
416 [44] Díaz R, Téllez-Téllez M, Bibbins-Martínez MD, Sánchez C, Díaz-Godínez G, Soriano-Santos GJ. Influence of initial
417 pH of the growing medium on the activity, production and expression profiles of laccases produced by *Pleurotus ostreatus*
418 in submerged fermentation. *Elect. J. Biotech. Bol.* 2013;16(4):1-13. <http://dx.doi.org/10.2225/vol16-issue4-fulltext-6>
419 [45] Córdoba-Sosa G, Torres JL, Ahuactzin-Pérez M, Díaz-Godínez G, Díaz R, Sánchez C. Growth of *Pleurotus ostreatus*
420 ATCC 3526 in different concentrations of di(2-ethylhexyl) phthalate in submerged fermentation. *J. Chem. Biol. Phys.*
421 *Scien.* 2014;4(5):96-103. doi=10.1.1.1075.6451
422 [46] Cedeño Domínguez JC, J.G. Merino Cordero JG. Valoración in vitro de polietileno de
423 bajadensidad mediante hongos filamentosos aislados del relleno sanitario de Pichacay. Carrera de Ingeniería Ambiental.
424 Universidad Politécnica Salesiana. Sede Cuenca. Tesis Licenciatura. Ingeniería Ambiental. 2020. Cuenca,
425 Ecuador. <http://dspace.ups.edu.ec/handle/123456789/18821>. Español
426 [47] Ramírez Cuadro NE, Teherán JA. Potencial tolerante y de biodegradación del hongo de
427 podredumbre blanca en lantanas usadas. Facultad de Ingeniería, Arquitectura y Diseño. Cartagena. Tesis Licenciatura.
428 Ingeniería Química. 2017. Cartagena, Colombia. bibliotecadigital.usbcali.edu.co. Español
429 [48] Arancibia Cortes VE. Caracterización de *Alternaria alternata* aislada de contenedores residuales urbanos y
430 su potencial uso en la degradación de 6 polímeros de Importancia Ambiental. Universidad Santo Tomás. Tesis para optar
431 al título profesional de Tecnólogo Médico Mención Laboratorio Clínico, Hematología y Banco de Sangre. 2014. Viña del
432 Mar, Chile. DOI: 10.13140/RG.2.1.3310.0884. Español
433 [49] González-Márquez A, Loera-Corral O, Viniegra-González G, Sánchez C. Production of cutinolytic esterase by
434 *Fusarium culmorum* grown at different apple cutin concentrations in submerged fermentation. *Mex. J. Biotech.* 2019;4(4):
435 50-64. <https://doi.org/10.29267/mxjb.2019.4.4.50>
436 [50] Russell JR, Huang J, Anand P, Kucera K, Sandoval AG, Dantzer KW. Biodegradation of polyester polyurethane by
437 endophytic fungi. *Appl. Environ. Microb.* 2011;77(17): 6076-6084. DOI: <https://doi.org/10.1128/AEM.00521-11>
438 [51] Butron Pizano SB. Capacidad de biodegradación de *Pseudomonas aeruginosa* frente al polietileno de bajadensidad.
439 Escuela de Posgrado. Universidad Nacional del Altiplano. Tesis Doctoral. Ciencia, Tecnología y Medio Ambiente. 2020.
440 Puno, Perú. <http://repositorio.unap.edu.pe/handle/UNAP/13475>. Español
441 [52] Johnnie DA, Isaac R, Prabha ML. Bio efficacy assay of laccase isolated and characterized from *Trichoderma viride* in
442 biodegradation of low density polyethylene (LDPE) and textile industrial effluents dyes. *J. Pure Appl. Microbiol.*
443 2021;15(1):410-420. DOI:10.22207/JPAM.15.1.38
444 [53] Ocegüera-Cervantes A, Carrillo-García A, López N, Bolaños-Núñez S, Cruz-Gómez MJ, Wachter C. Characterization
445 of the Polyurethanolytic Activity of Two *Alicyclophillus* sp. Strains Able to Degrade Polyurethane and N-Methylpyrrolidone.
446 *Appl. Environ. Microbiol.* 2007;73(19): 6214-6223. doi: 10.1128/AEM.01230-07
447 [54] Shah A, Hasan F, Akhter JI, Hameed A, Ahmed A. Degradation of polyurethane by novel bacterial consortium
448 isolated from soil. *Ann. Microb.* 2008;58(3):381-386. <https://doi.org/10.1007/BF03175532>
449 [55] Ferrer-Parra L, López-Nicolás DI, Martínez-Castillo R, Montiel-Cina JP, Morales-Hernández AR, Ocaña-Romo E.
450 Caracterización parcial de esterasas de *Fusarium culmorum* crecido en presencia de di(2-etil hexilftalato)
451 en fermentación sólida y sumergida. *Mex. J. Biotechnol.* 2018; 3(1):82-94. <https://doi.org/10.29267/mxjb.2018.3.1.84>.
452 Español

453 [56]Satti SM, Shah Z, Luqman A, Hasan F, Osman M, Ali Shah A. Biodegradation of poly(3-hydroxybutirate) and poly(3-
454 hydroxybutirate-co-3-hydroxyvalerate) by newly isolated *Penicillium oxalicum* SS2 in soil microcosms and partial
455 characterization of extracellular depolymerase. *Curr. Microbiol.* 2020;77:1622-1636.doi: 10.1007/s00284-020-01968-7

UNDR PEER REVIEW