

**EFFECT OF BIODIGESTER EFFLUENT ON CARBOHYDRATE METABOLIZING  
ENZYMES, GLUTATHIONE AND GLUTATHIONE-RELATED ENZYME IN THE  
AFRICAN CATFISH (*Clarias Gariepinus*)**

**Abstract**

Research on the use of biodigester effluent as a feed source in fish farming have been previously reported, however, its effect on the metabolic activities of fish, which might inform its safety and administration guideline in fish farming is poorly understood. The aim of the present study was to determine the effect of pig dung biodigester effluent (PDBE) on the activities of lactate dehydrogenase (LDH), glucose-6-phosphate dehydrogenase (G-6-PDH), glutathione peroxidase (GPx), concentration of reduced glutathione (GSH) and total protein in the African Catfish (*Clarias gariepinus*). Fifteen catfish ( $120 \pm 10$  g) were purchased, acclimatized to laboratory conditions and grouped into three. Group I (control): exposed to freshwater; group two: exposed to raw PDBE; and group three: exposed to 1:10 dilution of PDBE. The experiment lasted for 24 hours after which the serum of the fish was used for metabolic analysis. LDH and G-6-PDH activities and total protein concentration were measured using Randox kits following the manufacturer's protocol while reduced glutathione concentration was measured using Ellman's method and glutathione peroxidase activity was also measured using Ellman's method after protein precipitation. Compared with control, a significant ( $P < 0.001$ ) increase in serum total protein concentration and the activities of lactate dehydrogenase, glucose-6-phosphate dehydrogenase, and glutathione peroxidase as well as a decrease in reduced glutathione concentration was found in group two exposed to raw PDB effluent. However, the metabolic parameters of group three catfish were nearly similar to that of control. Findings of the present study suggests that before using raw pig dung biodigester effluent for fish farming it should be

diluted using a 1:10 dilution. The present study has provided a comprehensive data for the development of the right protocol and guideline for the administration of biodigester effluent in catfish farming. It has also provided biomarkers which can be measured for the evaluation of biodigester effluent toxicity. Findings have implication for the production of low-cost fish feed for catfish.

**Keywords:** Effluent, G-6PDH, LDH, GSH, GPx, total protein.

## **Background**

*Clarias gariepinus*, also known as African sharp tooth catfish is a species of the air breathing catfish which grows very fast and multiply quickly [1]. It is indigenous to the inland freshwaters of Africa that is why it is called the African catfish [2]. Nigeria is the world's largest commercial producer of the African catfish. Given the continuous rise in the demand for catfish, catfish farming is currently a very lucrative business in Nigeria [3, 4]. Catfish production occurs in several communities in Bayelsa State [4]. Bayelsa is a State in Nigeria, which lies between River State and Delta State [5] and is part of the largest wetland in Africa, the Niger Delta [6]. Catfish ponds and sales are very common in Bayelsa State [4, 7]. Aside from *Clarias gariepinus*, other species of catfish found in Nigeria are; *Heterobranchus bidorsalis*, a hybrid of *Clarias gariepinus* and *Heterobranchus bidorsalis* known as *Heteroclarias*, and *Clarias nigro-digitatus* [8].

The high cost of fish feed is a major constraint to catfish production [4, 7]. A previous study found that fish feed accounted for 77.4% of the total cost of catfish production [9]. The conventional pelleted floating feeds for catfish consist of oil seed meal such as groundnut cake, Soybean cake, soybean flour [8], cotton seed meal and canola meal [10]. Sometimes unconventional animal by-products that meet the nutrient requirements of catfish are also used as feed for catfish [10].

Active research on fish feed formulations with high quality and low cost is on-going [11, 12]. A previous research showed that fish residing close to wastewater effluent discharges tend to have more weight than those of control [13]. Biodigester effluent is a type of wastewater that is released from a biodigester during anaerobic digestion of biodegradable wastes [14]. A previous study showed that bio-digester effluents contained a good amount of nitrogen, potassium, phosphates, calcium, magnesium and sodium but with *E. coli* presence which was removed upon treatment [15]. This suggests that wastewater effluent contains food and nutrients which are needed for fish growth [13]. Other previous studies have reported the use of biodigester effluent as a rich source of nutrient for plant growth [16, 17] and as fertilizer for fish growth [18, 19]. Although studies have showed that biodigester effluent can be used as fish feed, however, evaluation of its effect on key pollution biomarkers in catfish is limited in literature. Concerns have been raised over the safety of biodigester effluent as fish feed [20,21]. Concern has also been raised over environmental pollution during its application [19].

Common metabolic markers of pollution and toxicity in fish and other lower animals which have been reported in previous studies are Glucose-6-phosphate dehydrogenase (G-6-PDH), total protein, glutathione peroxidase (GPX), reduced glutathione (GSH) and lactate dehydrogenase (LDH) and other antioxidant enzymes [22 - 27]. Therefore, in order to determine the safety protocol and guideline underlying the use of biodigester effluent in catfish farming, the present study investigated the effect of pig dung biodigester effluent (PDBE) on the activities of lactate dehydrogenase, glucose-6-phosphate dehydrogenase, glutathione peroxidase, reduced glutathione and total protein concentration in African Catfish (*Clarias Gariepinus*).

## **2. Materials and methods**

## **2.1 Study location**

This research was conducted in the University of Africa Toru –Orua, Sagbama Local Government Area (LGA), Bayelsa State, Nigeria. The State is situated on the core of Niger Delta, Nigeria [28] on longitude 6.06990° E latitude and 4.77190° N [29]. Toru-Orua is a swampy environment that is well watered by the freshwater from ForcadosRiver, surrounding creeks and lakes that flows across it [30].

## **2.2 Animals, biowaste and study design**

Fishing is a common occurrence in Bayelsa State [31] and fifteen catfish ( $120 \pm 10$  g) were purchased from one of live catfish traders in Toru – Orua community. They were acclimatized to laboratory conditions for four hours in freshwater [32]. Effluent from pig dung biodigester was obtained from the Biogas Production and Research Centre in the University of Africa Toru – Orua, Bayelsa State. The catfish were grouped into three. Group I (control water): exposed to 20 litres of freshwater; group two exposed to 20 litres of raw effluent from pig dung biodigester; group three: exposed to diluted effluent from pig dung biodigester. Freshwater was used for dilution. International guidelines for the use of fishes in research was followed for the care and handling of the experimental catfish used in the present study [33].

## **2.3 Acute toxicity test**

The short-term adverse effects of exposing catfish to two different concentrations of pig dung effluent was evaluated for a period of 24 hours [34, 13]. At the end of the experiment, blood sample was collected from the fish. The universal ethical recommendations for taking blood from catfish was followed [35]. Blood was drawn into plain serum bottles for serum preparation [26].

## 2.4 Determination of total protein (TP) concentration

This was carried out using the manufacturer's protocol for Randox total protein kit based on the Biuret method, with bovine serum albumin as the protein standard [36].

**Principle:** Cupric ions, in an alkaline medium, interact with protein peptide bonds resulting in the formation of a blue coloured complex, which exhibits maximum absorbance between wavelengths of 530-570 nm [36].

**Procedure:** Following the Randox kit protocol, 1 ml of reagent R1 [containing Sodium hydroxide (100 mmol/l), sodium-potassium tartrate (16 mmol/l), potassium iodide (15 mmol/l) and copper II sulphate (6 mmol/l)] was added to 0.02 ml of serum sample. The mixture was incubated at 25°C and the absorbance was measured against reagent blank at a wavelength of 546 nm. The total protein concentration was expressed as mg/dl.

## 2.5. Estimation of serum lactate dehydrogenase activity

Lactate dehydrogenase (LDH) activity was carried out following the manufacturer's protocol for Randox LDH Kit using UV method [37].

**Principle:**  $\text{Pyruvate} + \text{NADH} + \text{H}^+ \xrightarrow{\text{LDH}} \text{L-Lactate} + \text{NAD}^+$

The activity of the LDH that leaked from tissues into the blood during tissue damage was measured in the present study. The reaction principle is based on the catalytic activity of serum LDH which leads to the reversible reduction of pyruvate to L-lactate mediated by the hydrogen donor, NADH. The decrease in NADH, was harnessed as a basis for the measurement of LDH activity. NADH in solution produces a significant absorbance peak at 340 nm, while NAD<sup>+</sup> has virtually no absorbance at this wavelength. This difference was the basis by which LDH activity was monitored. For the calculation of LDH activity, the rate of absorbance change per minute at

340 nm due to the reduction of NADH was multiplied by 9683 (a factor given in the manufacturer's protocol for Randox LDHKit) and the calculated activity was expressed as IU/L/min [37].

**Protocol:** Following the Randox kit protocol, 0.02 ml of serum sample was pipetted into a test tube followed by 1 ml of Randox reagent. The Randox reagent contained phosphate buffer (50 mmol/l, pH 7.5), pyruvate (substrate 0.6 mmol/l) and NADH (0.18 mmol/l). The serum sample and Randox reagent were thoroughly mixed together and absorbance was read at 340 nm at 30 sec, 1, 2 and 3 min intervals. The activity of LDH was calculated by multiplying the change in absorbance per min ( $\Delta A$  340 nm/min) with 9683 (a factor given in the manufacturer's protocol for Randox LDHKit). The calculated activity was expressed as IU/L/min.

## 2.6 Estimation of Glucose-6-Phosphate Dehydrogenase (G-6-PDH) Activity

This was carried out following the manufacturer's protocol for Randox G-6-PDHKit based on UV method [38].

**Principle:** The activity of G-6-PDH in the serum was determined by the measurement of the absorbance change at 340 nm due to the reduction of  $NADP^+$ . This enzymatic dehydrogenase reaction takes advantage of the ability of the reduced form (NADPH), to absorb light at a wavelength of 340 nm while the oxidized form ( $NADP^+$ ) do not [38].



**Protocol:** Following the manufacturer's protocol for Randox kit, 1 ml of reagent R1 containing triethanolamine buffer (31.7 mmol/l, pH 7.6) and EDTA (3.2 mmol/l) was pipetted into a test tube, followed by 0.03 ml of reagent R2 containing NADP (0.34 mmol/l) reconstituted with 2 ml of distilled water and 0.015 ml of serum sample. The mixture was mixed thoroughly and

incubated for 5 min at 37<sup>0</sup>C. Reagent 3 containing glucose-6-phosphate dehydrogenase substrate (0.34 mmol/l) was reconstituted with 2 ml of redistilled water and was added to the mixture. The whole content of mixture was mixed thoroughly and the absorbance was taken at 0 sec 1, 2 and 3 minutes.

Activity of G-6-PDH (U/L/min) = 33650 (factor) x change in absorbance at 340 nm/min.

## **2.7 Determination of Reduced Glutathione (GSH) Concentration**

The method originally described by Ellman [39] was used in estimating the level of reduced glutathione (GSH).

**Principle:** The reduced form of glutathione comprises in most instances the bulk of cellular non-protein sulfhydryl groups. The principle of reaction is therefore based upon the oxidation of glutathione by the sulfhydryl reagent 5,5'-dithio-bis(2-nitrobenzoic acid) (DTNB) resulting in the formation of a relatively stable yellow chromogenic derivative called 5'-thio-2-nitrobenzoic acid (TNB), measured at a wavelength of 412 nm. The sulfhydryl reagent is the Ellman`s reagent.

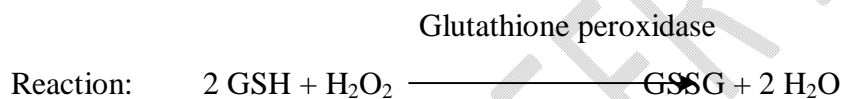
**Procedure:** Liver homogenate (0.2 ml) was added to 1.8 ml of distilled water and 3 ml of protein precipitating solution (2 ml of 5% TCA). The mixture was mixed together and allowed to stand for approximately 5 min, after which it was centrifuged and filtered. Exactly 1ml of the supernatant was added to 3.0 ml of 0.2 M phosphate buffer (pH 8.0) and 0.5 ml of Ellman`s reagent (19.8 mg of 5', 5'-dithiobis-2-nitrobenzoic acid, DTNB in 100 ml of 0.1% sodium nitrate) was added and the absorbance was read at 412 nm against a reagent blank. Reduced GSH concentration is proportional to the absorbance of the final mixture at 412 nm wavelength. GSH concentration was extrapolated from the standard curve for reduced GSH [39].

## 2.8 Evaluation of glutathione peroxidase activity

Glutathione peroxidase activity was determined according to the method of Rotruck *et al.*, [40].

**Principle:** Glutathione peroxidase catalyzes the oxidation of hydrogen peroxide ( $\text{H}_2\text{O}_2$ ) using reduced glutathione. The reduced glutathione is oxidized in the process. Glutathione peroxidase activity is arrested by TCA (Trichloroacetic acid) and the remaining glutathione in the reaction mixture is determined by the method originally described by Ellman [39]. In this protocol, thiols present in the remaining glutathione in the reaction mixture interact with 5, 5'-dithiobis-(2-nitrobenzoic acid) (DTNB) to form a highly yellow coloured anion with maximum peak at 412nm. Glutathione peroxidase activity is proportional to the amount of glutathione that was consumed in the reaction.

### Equation of reaction



Abbreviations used: GSH = Reduced glutathione; GSSG = oxidized glutathione

**Procedure:** The reaction mixture contained 0.2 ml of 0.4 M Tris-HCl buffer, pH 7.0, 9.1 ml of 10 mM sodium azide, 0.2 ml of liver homogenate, 0.2 ml of 4 mM glutathione and 0.1 ml of 0.2 mM hydrogen peroxide. The reaction mixture was incubated at 37°C for 10 min and was arrested by the addition 0.4 ml of 10% TCA, followed by centrifugation at 3000 rpm for 5 min. The supernatant was assayed for glutathione content by using Ellman's reagent. Briefly, to 1ml of the supernatant was added 0.5 ml of Ellmans reagent (19.8 mg of 5', 5'-dithiobis-2-nitrobenzoic acid, DTNB in 100 ml of 0.1% sodium nitrate) and 3.0 ml of 0.2 M phosphate buffer (pH 8.0) and the absorbance was read at 412 nm against a reagent blank. Glutathione peroxidase activity was obtained by plotting a GSH standard curve and the concentration of the remaining GSH in

the reaction mixture was obtained from the standard calibration curve by extrapolation from the curve.

**Calculation:** GSH consumed = initial GSH concentration – remaining GSH in the mixture.

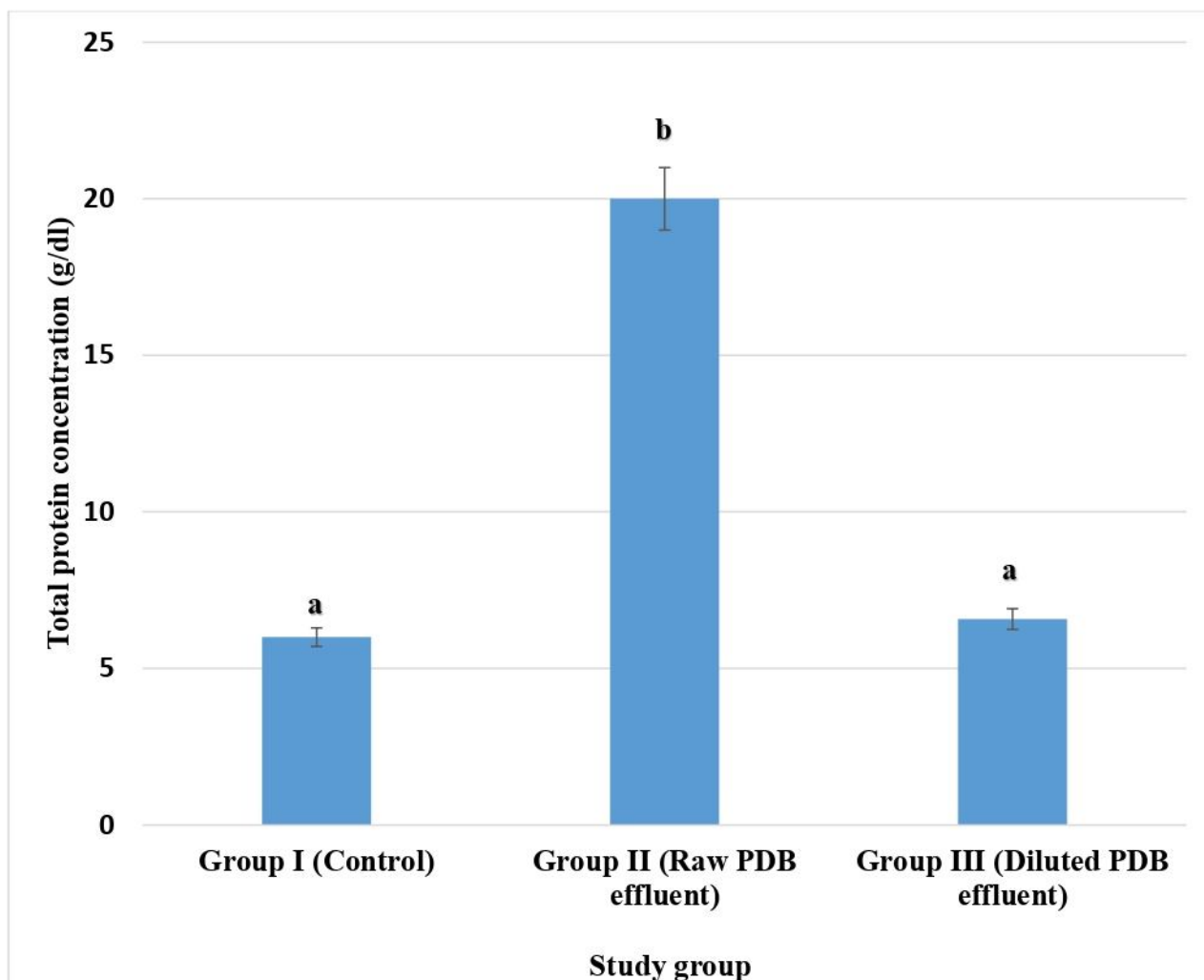
## 2.9 Statistical analysis

IBM SPSS software version 24 was used for data analysis. One-way ANOVA was used to test for significant difference between groups and results were presented as mean  $\pm$  standard error of mean of five replicates. *P*-value was set at 0.05 [41, 42].

## 3.0 Results

### 3.1 Effect of PDB effluent on serum total protein concentration in African catfish (*Clarias gariepinus*)

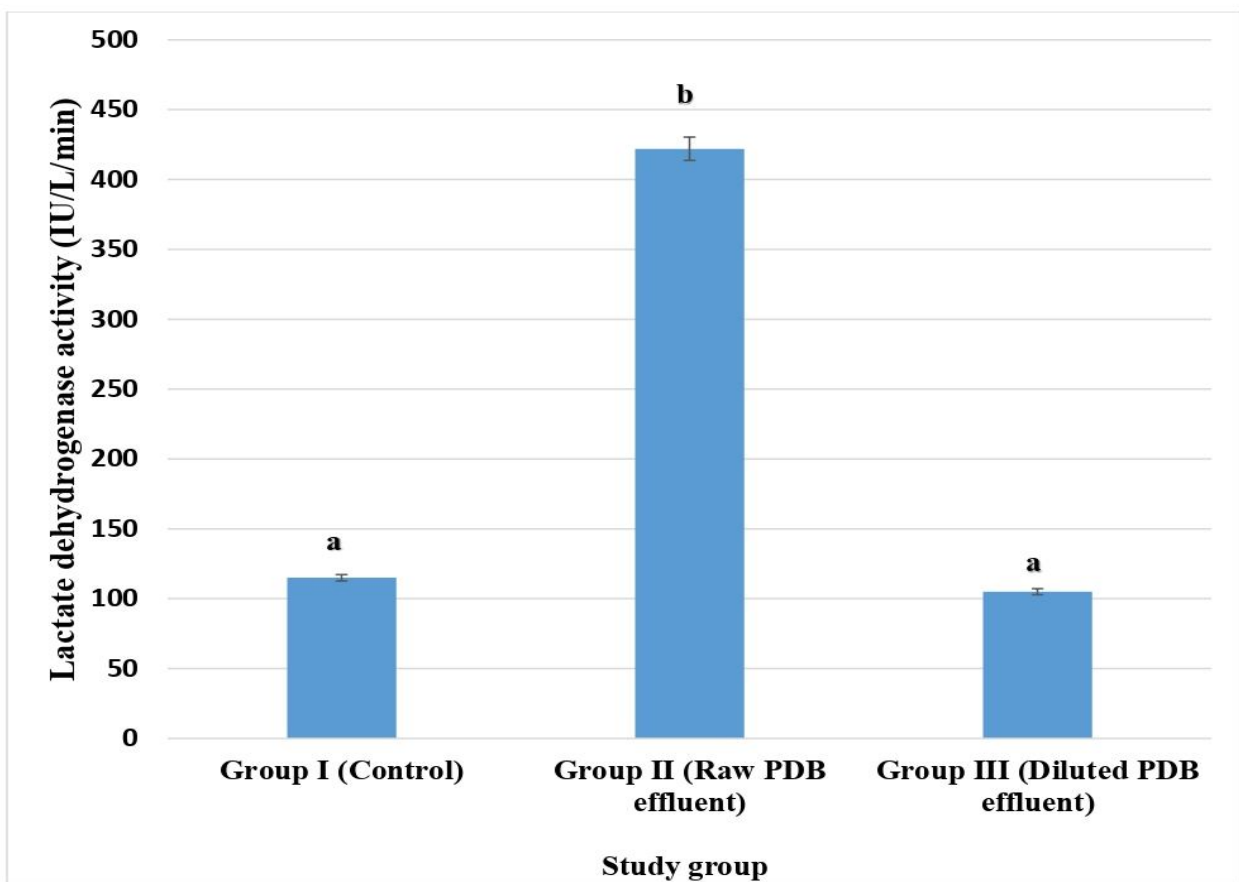
The serum total protein concentration of African catfish (*Clarias gariepinus*) exposed to PDB effluent is presented in Figure 1. Compared with control, a significant ( $P < 0.001$ ) increase in total protein concentration was found in the serum of group 2 catfish exposed to raw PDB effluent (control:  $6.00 \pm 0.12$  g/dl versus group II:  $20.0 \pm 0.10$  g/dl), while no significant difference was found in group 3 fish compared with control (control:  $6.00 \pm 0.12$  g/dl versus group III:  $6.86 \pm 0.02$  g/dl).



**Figure 1. Concentration of serum total protein in the serum of African catfish (*Clarias gariepinus*) exposed to PDB effluent.** Result presented as mean± standard error of mean (SEM); bars with different superscripts are significantly different ( $P < 0.05$ ) while bars with the same superscript are not significantly different. PDB: pig dung biodigester. Group 1 fish (control): exposed to freshwater. Group 2: exposed to raw PDB effluent and group 3: exposed to diluted PDB effluent (1:10 dilution).

### 3.2 Effect of pig dung biodigester effluent on the activity of serum lactate dehydrogenase in African catfish (*Clarias gariepinus*)

As shown in Figure 2, compared with control, a significant ( $P < 0.001$ ) increase in the activity of LDH was found in group 2 catfish exposed to raw PDB effluent (control:  $115.00 \pm 1.57$  IU/L versus group 2:  $422.00 \pm 5.64$  IU/L). On the contrary, no significant difference in the activity of LDH was found in group 3 catfish ( $105.00 \pm 5.18$  IU/L) compared with control.

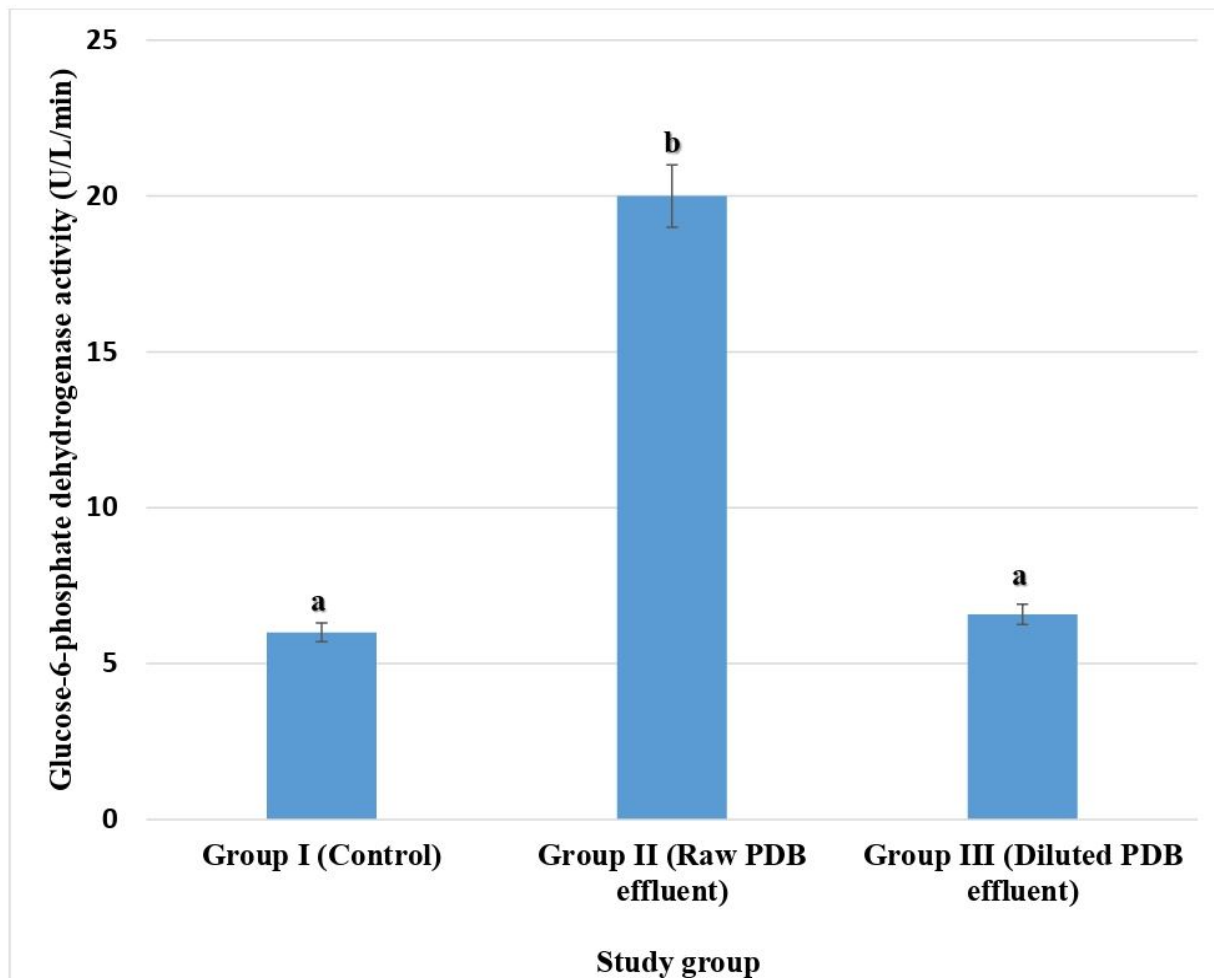


**Figure 2** Effect of PDB effluent on serum lactate dehydrogenase (LDH) activity in African catfish (*Clarias gariepinus*). Result presented as mean  $\pm$  standard error of mean (SEM). Bars with different superscripts are significantly different ( $P < 0.05$ ) while bars with the same superscript are not significantly different. PDB: pig dung biodigester. Group 1 (control): exposed

to freshwater. Group 2: exposed to raw PDB effluent and group 3: exposed to diluted PDB effluent (1:10 dilution).

### **3.3 Effect of pig dung biodigester effluent on the activity of serum glucose-6-phosphate dehydrogenase in African catfish (*Clarias gariepinus*)**

Results for the effect of PDB effluent on the enzyme activity of glucose-6-phosphate dehydrogenase (G-6-PDH) in the serum of catfish (*Clarias gariepinus*) is presented in Figure 3. As shown in Figure 3, compared with control a significant ( $P < 0.001$ ) increase in the activity of G-6-PDH was found in the serum of group II (control:  $8.00 \pm 0.40$  U/g versus group II:  $32.00 \pm 0.73$  U/g;  $P < 0.001$ ), while the activity of G-6-PDH in group III catfish exposed to diluted PDB effluent was found to be similar to that of control with no significant difference (control:  $8.00 \pm 0.40$  U/g versus group III:  $7.00 \pm 0.40$  U/g;  $P > 0.05$ ).



**Figure 3** Effect of PDB effluent on serum glucose-6-phosphate dehydrogenase (G-6-PDH) activity in African catfish (*Clarias gariepinus*). Result presented as mean $\pm$  standard error of mean (SEM). Bars with different superscripts are significantly different ( $P < 0.05$ ) while bars with the same superscript are not significantly different. PDB: pig dung biodigester. Group 1 (control) exposed to freshwater. Group 2: exposed to raw PDB effluent and group 3: exposed to diluted PDB effluent (1:10 dilution).

### 3.4 Effect of pig dung biodigester effluent on reduced glutathione concentration in the serum of African catfish (*Clarias gariepinus*)

Results for the effect of PDB effluent on the concentration of reduced glutathione in the serum of catfish (*Clarias gariepinus*) is presented in Figure 4. As shown in Figure 4, compared with control a significant ( $P < 0.001$ ) decrease in the concentration of reduced glutathione was found in group II. However, significant ( $P < 0.001$ ) elevation in the concentration of reduced glutathione was found in group III compared with control.

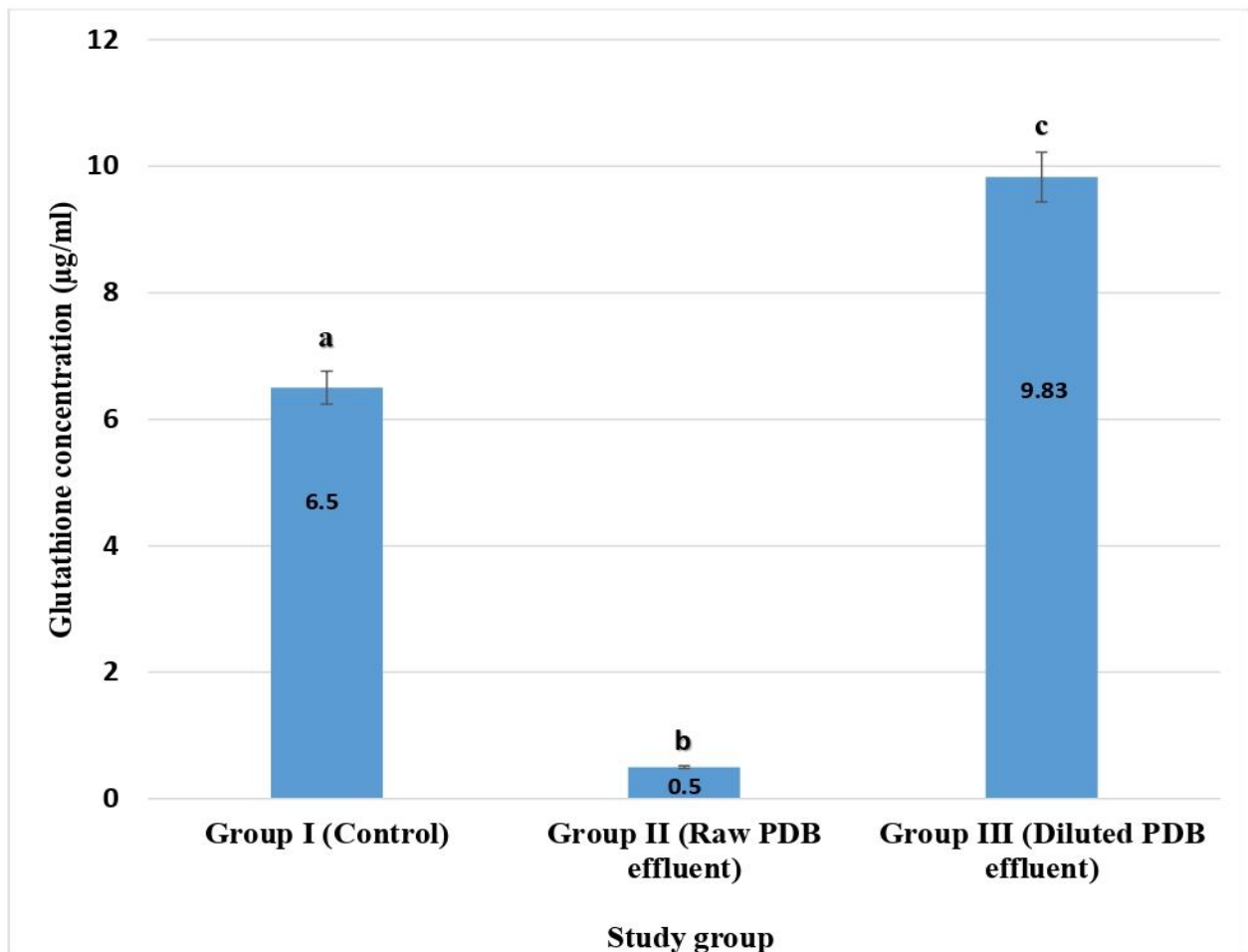
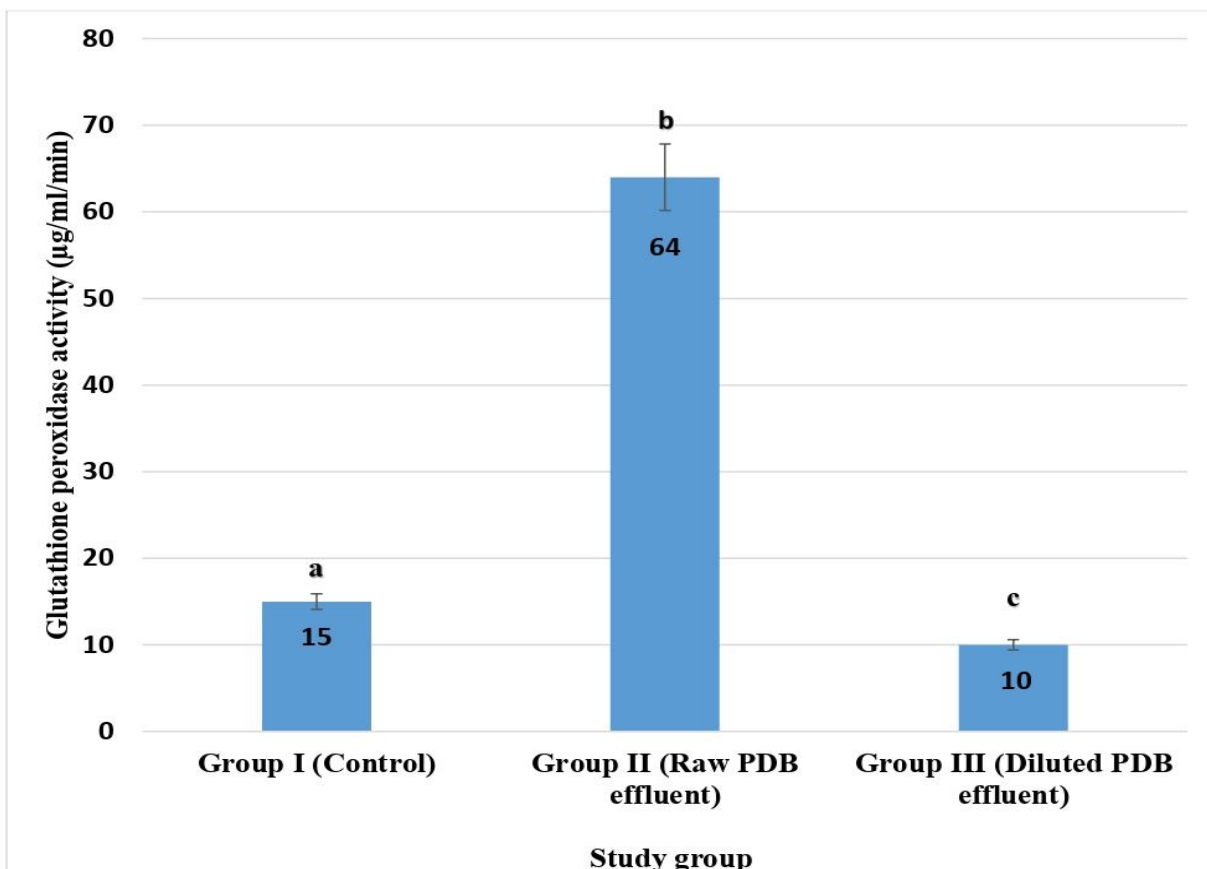


Figure 4 Effect of PDB effluent on reduced glutathione in the serum of African catfish (*Clarias gariepinus*). Result presented as mean  $\pm$  standard error of mean (SEM). Bars with

different superscripts are significantly different ( $P < 0.001$ ) PDB: pig dung biodigester. Group 1 (control): catfish exposed to freshwater. Group 2: catfish exposed to raw PDB effluent, and group 3: catfish exposed to diluted PDB effluent (1:10 dilution).

### 3.5 Effect of pig dung biodigester effluent on serum glutathione peroxidase activity in African catfish (*Clarias gariepinus*)

As shown in Figure 5, compared with control, the activity of serum glutathione peroxidase (GPx) was found to be significantly higher in group II catfish exposed to raw PDB effluent than in control. However, the enzyme activity of GPx decreased significantly ( $P < 0.001$ ) lower than that of control following a 1:10 dilution of PDB effluent.



**Figure 5 Effect of PDB effluent on the activity of glutathione peroxidase in the serum of African catfish (*Clarias gariepinus*).** Results presented as mean $\pm$  standard error of mean (SEM). Bars with different superscripts are significantly different ( $P < 0.001$ ). PDB: pig dung biodigester. Group 1 (control): catfish exposed to freshwater sample. Group 2: catfish exposed to raw PDB effluent, and group 3: catfish exposed to diluted PDB effluent (1:10 dilution).

## Discussion

Comparative evaluation of the biomarkers of pollution in catfish exposed to raw PDB effluent and catfish exposed to diluted PDB effluent was carried out in the present study. Both LDH and G-6-PDH are carbohydrate metabolizing enzyme that serve as biomarkers for pollution and toxicity [25]. Given that the liver is one of the richest sources of LDH [37], the increase in serum LDH activity beyond normal serum level which was found in the present study in group two after exposure to pig dung biodigester effluent suggests that raw PDBE induced excessive damage and destruction to the liver cells of catfish which may have led to increased permeability of liver cells and leakage of LDH as well as other parameters into the blood. However this was not found in group 3 catfish which were placed in diluted PDBE

Findings of the present study corroborates findings of a previous study where an increase in lactate dehydrogenase was found in the freshwater fish called *Channa punctatus* following exposure to increasing concentration of tannery waste water [43]. Increase in the activity of LDH was also found in the liver and muscle cells of tilapia fish (*Oreochromis mossambicus*) following exposure to fumaronitrile, a persistent organic pollutant of industrial waste water [27]. Significant changes in the activity of LDH was also previously found in the armored catfish,

*Rhinelepis strigosain* (Siluriformes, Loricariidae) after exposure to hypoxia and temperature variation compared with the unexposed group [44].

A previous study found that exposure of fish to cadmium and lead also resulted in a significant decrease in LDH and G-6-PDH activities [45]. Furthermore, another previous study also showed a reduction in the activity of glucose-6-phosphate dehydrogenase in the erythrocytes of dogfish and 5 other fish species from Black Sea after exposure to stress [46]. These findings are consistent with the findings of the present for group 2. However the reverse was found for group III catfish placed in diluted PDBE in the present study.

Enzymatic and non-enzymatic antioxidants serve as important biological defense against environmental oxidative stress and have been reportedly used as biomarkers of pollution in several animal studies [47]. Comparative evaluation of reduced glutathione levels between catfish exposed to raw PDB effluent and catfish exposed to diluted PDB effluent was also carried out in the present study. The present study found a decrease in glutathione concentration and an increase in the antioxidant enzyme activity of catfish exposed to raw PDB effluent. This is in agreement with the findings of previous studies where a decrease in glutathione level was found in the freshwater snail called *Bellamyapurificata* after exposure to landfill leachate effluent and bisphenol A [48]. The same was observed in the fish, *Channa punctatus* after exposure to a thermal power plant effluent in a previous study [49]. Elevation of glutathione peroxidase activity has also been reported in a previous study where the effect of waste water on fish was evaluated [50]. The present study suggests that the induction of glutathione peroxidase activity was caused by pollutants present in PDBE. GSH and GPx may be considered as potential specific biomarkers for PDBE pollution.

The present study found that the administration of 1:10 dilution of PDB effluent to group III catfish restored the altered metabolic parameters in group II to that of control and better than that of control for GSH and GPx. This suggests that the concentration of PDBE should be taken into consideration when using it as feed source or supplement in fish farming. The present study is quite novel because while previous studies mainly focused on only the effect of raw wastewater on fish, the present study focused on the effect of both the raw waste water and the diluted wastewater in order to determine the proper guideline for the application of PDBE for fish farming.

Aside from providing a guideline for PDBE administration in fish farming, the present study has also showed that when the levels of pollutants exceed safe limits, a reduction in the overall health of fish can occur. These impacts can also adversely affect human health (Mustafa, 2024).

## **Conclusion**

The knowledge provided in the present study might aid the understanding of PDBE utilization in fish farming. It will also be a significant resource for fish farm managers and authorities to improve on the formulation of low cost fish feeds for fish farming. It can be concluded from the present study that although fishes had reduced metabolic activity after been exposed to a high concentration of raw PDBE, however, the metabolic activities of catfish greatly improved in the catfish exposed to diluted PDBE. This implies that at a high concentration of PDB effluent impairs metabolism in catfish, while at a low concentration of PDB effluent enhances metabolism in catfish. The result of the present study also implies that total protein, G-6-PDH, LDH, GSH and GPx will be valuable tools in biological monitoring of PDBE pollution. In

addition, the present study has provided comprehensive data for the development of guidelines for the administration of biodigester effluent in aquaculture. It has also provided biomarkers to be measured for the evaluation of biodigester effluent toxicity. Finally, the present results suggests that a 1:10 dilution of PDBE might be a very effective practice for the detoxification of PDBE for catfish farming.

## References

1. Zhang F, Wan W, Li Y, Wang B, Shao Y, Di X, Zhang H, Cai W, Wei Y, Ma X. (2024). Construction of a full-length transcriptome resource for the African sharptooth catfish (*Clarias gariepinus*), a prototypical air-breathing fish, based on isoform sequencing (Iso-Seq). *Gene*, 930:doi.org/10.1016/j.gene.2024.148802.
2. Mbanga B, van Dyk C, Maina JN. (2018). Morphometric and morphological study of the respiratory organs of the bimodally-breathing African sharptooth catfish (*Clarias gariepinus*): Burchell (1822). *Zoology*, 130:6-18.
3. Food and Agriculture Organization of the United Nations (2024). Fish4ACP: Unlocking the potential of sustainable fisheries and aquaculture in Africa, the Caribbean and the Pacific. <https://www.fao.org/in-action/fish-4-acp/where-we-work/africa/nigeria/en/>.
4. Onuwa G, Mailumo SS, Oyewole OS. (2023). Empirical analysis of catfish productivity among smallholders in Ekeremor, Bayelsa State, Nigeria. *IJANS*, 15(3):257–263.
5. Ogbole FA & Sanugba EE. (2023). Distribution of body mass index and abdominal obesity in Bayelsa State with associated interleukin-2 gene expression. *AJBGMB*, 14(4):1–10.
6. Ogbole FA, Igwe CU, Onuoha HC & Nzebude CP. (2023). Evaluating the prevalence of malaria parasite infection among adults in wetlands using nested PCR and high resolution melting analysis. *AJBGMB*, 14(4):53–63.
7. Kainga PE, Okpukpara BC and Morgan CN. (2019). Technical efficiency of catfish (*Clarias gariepinus*) production in Bayelsa State, Nigeria: a stochastic approach. *AJATE*, 8(2):61-70.
8. Adewumi AA and Olaleye VF. (2011). Catfish culture in Nigeria: Progress, prospects and problems. *AJEST*, 8(2):001-005.

9. Adeosun OM, Olaoye OJ, Ojebiyi WG, Agarawu OL& Adeosun, FI. (2024). Profitability analysis of catfish production in Odogbolu Local Government Area of Ogun State, Nigeria. *NJAAT*, 4(2):187–202.
10. Fish site (2013). Fish ingredients and feed for Channel catfish. Retrieve from: <https://the fishsite.com>articles>feed ingredient-and fish feed>.
11. Yahaya MS, Biu MA and Abdulsalam S. (2021). Formulation, production and evaluation of floating catfish feed. *NAJ*, 46(2):168 – 174.
12. Olapade OJ and Saboleh P. (2022). Development of fish feeds for African catfish (*Clarias gariepinus* Burchell 1822) farming in Sierra Leone, West Africa. *IJFA*, 14(2):15-21.
13. Andrade Muñoz AS, Di Prinzio CY, Assef YA, Kutschker AM, Alday, GL et al (2023). Implications of wastewater discharges on environmental features and fish communities in an urban river; Springer; *Urban Ecosyst*, 26(3):779-791.
14. Ogbole FA, Ogbuta AA, Okagbue HI. (2024). Characterization of fresh biowastes for biogas production and environmental health in Niger Delta, Nigeria. *BMC Environ Sci* 1(8):1-13.
15. Barzallo-Bravo LA, Carrera-Villacrés D, Vargas-Verdesoto RE, Karina PL, Correoso M, Gavilanes-Quishpi AP. (2019). Bio-digestion and post-treatment of effluents by bio-fermentation, an opportunity for energy uses and generation of organic fertilizers from bovine manure. *Int J Recycl Org Waste Agricult* 8:431–438.
16. Rosety-Rodriguez M, Ordonez F, Rosety I, Rosety J, Rosery M. (2005) Erythrocyte antioxidant enzymes of gilthead as early-warning bio-indicators of oxidative stress induced by malathion. *Haema*, 8:237-240.
17. Southavong S, Khammingsavath K, Vyraphet P, Preston TR. (2012). Effect of effluent-treated biochar and biodigester effluent on growth of maize (*Zea mays*) and on soil physical properties. *LRRD*, 24:104. Retrieved from <http://www.lrrd.org/lrrd24/6/24104.htm>
18. Ngan NVC, Be NV, Du NX. (2017). Study on fish growing in hapa conditions using bio-slurry from co-digestion process. *IJASRM*, 2(6):69 – 76.
19. Nhi NHY, Preston TR. (2011). The growth and economics of integrated culture of Tilapia (*Oreochromis niloticus*) and Common carp (*Ciprinus carpio*) in an indoor intensive system with earthworms as feed and in natural ponds fertilized with biodigester effluent and supplemented with duckweed. *LRRD*, 23(161). Retrieved from <http://www.lrrd.org/lrrd23/7/nhi23161.htm>.
20. Ogbole FA and Akemi CO. (2023). GC-MS analysis of biogas from pineapple peels and toxicological evaluation of generated effluent. *AJBGE*, 6(2):96-104.

21. Ogbole FA, Ebisintei P. (2024). Effect of biogas production effluent on oxidative stress, antioxidant enzymes and behavioural characteristics of African catfish (*Clarias gariepinus*). *IJRPR*, 5(8):2898-2905.
22. Tsouko E, Khan A, White M, Han JJ, Shi Y, Merchant FA, Sharpe MA, Xin L, Frigo DE. (2014). Regulation of the pentose phosphate pathway by an androgen receptor–mTOR-mediated mechanism and its role in prostate cancer cell growth. *Oncogenesis*, 3:e103. [doi.org/10.1038/oncsis.2014.18](https://doi.org/10.1038/oncsis.2014.18)
23. Zhang Y, Xu Y, Lu W, Li J, Yu S, Brown EJ, Stanger BZ, Rabinowitz JD, Yang X. (2022). G6PD-mediated increase in de novo NADP<sup>+</sup> biosynthesis promotes antioxidant defense and tumor metastasis. *Sci Adv*. 22;8(29):eabo0404. doi: 10.1126/sciadv.abo0404.
24. Sun L, Sun B, Zhang, Y, Chen, K. (2024). Kinetic properties of glucose 6-phosphate dehydrogenase and inhibition effects of several metal ions on enzymatic activity in vitro and cells. *Sci Rep* 14, 5806. <https://doi.org/10.1038/s41598-024-56503-6>
25. Ogbole FA, Crown OO, Olayeriju OS, Olaleye MT, Akindahunsi AA. (2019). Antidiabetic effect of methanolic extract of *Garcinia kola* leaves on streptozotocin-induced diabetic rats. *GSJ*, 7 (4):634 – 641.
26. Ogbole FA, Crown OO, Olayeriju OS, Olaleye MT, Akindahunsi AA.(2019). Hepatoprotective and antidyslipidemic effect of methanolic extract of *Garcinia kola* leaves on streptozotocin-induced diabetic rats. *IJEAST*, 4(4):1-5.
27. Chinnadurai K, Prema, P, Veeramanikandan V, Kumar KR, Nguyen V, Marraiki N, Zaghloul NSS, Balaji P. (2022). Toxicity evaluation and oxidative stress response of fumaronitrile, a persistent organic pollutant (POP) of industrial waste water on tilapia fish (*Oreochromis mossambicus*). *Environ Res*, 204:112030.
28. Ogbole FA, Moroyei BE. (2023). Upregulation of interleukin-2 among hypertensive subjects in Bayelsa State, Nigeria. *IJRR*, 10(6):557-565.
29. Ogbole FA. (2021). Urinalysis for Dehydration, Kidney Injury and Urinary Tract Infection Assessment in Rural Riverside, Bayelsa State, Nigeria. *IJISRT*, 6(12):805 – 810.
30. Ogbole FA, Oyelana O. (2020). Health Risk Assessment of the Drinking Water from Sagbama River, Bayelsa State, Niger Delta, Nigeria. *IJSER*, 11(9):1455 – 1460.
31. Ogbole FA, Harold BA. (2023). Association of undiagnosed pre-diabetes and type-2 diabetes mellitus with interleukin-2 mRNA expression among adults in Bayelsa State, Nigeria. *IJRR*, 10(5): 210-215.
32. Makaras T, Stankevičiūtė M, Šidagytė-Copilas E, Virbickas T, Razumienė J. (2021) Acclimation effect on fish behavioural characteristics: determination of appropriate acclimation period for different species. *J Fish Biol*. 99:502–512.

33. American Fisheries Society (2014). Guidelines for the Use of Fishes in Research. Bethesda, Maryland. Retrieved from <https://fisheries.org/docs/wp/Guidelines-for-Use-of-Fishes.pdf>.
34. United States Environmental Protection Agency (2024). Acute toxicity WET methods. Freshwater and marine organisms. Office of Water (4303T). 1200 Pennsylvania Avenue, N.W. Washington, DC 20460. Retrieved from: [epa.gov/cwa-methods/acute-toxicity-wet-methods#:~:text=The%20acute%20toxicity%20tests%20generally,from%2024%20to%2096%20hours](https://epa.gov/cwa-methods/acute-toxicity-wet-methods#:~:text=The%20acute%20toxicity%20tests%20generally,from%2024%20to%2096%20hours).
35. Argungu LA, Siraj SS, Christianus A, Amin MSN, Daud SK, Abubakar MS, Abubakar IA, Aliyu -Paiko M. (2017). A simple and rapid method for blood collection from walking catfish, *Clarias batrachus* (Linnaeus, 1758). *IJFS*, 16(3) 935 -944.
36. Weichselbaum TE. (1946). An accurate and rapid method for the determination of proteins in small amounts of blood serum and plasma. *Am J Clin Pathol*, 10:40-49.
37. Farhana A, Lappin SL. (2024). Biochemistry, Lactate Dehydrogenase. [Updated 2023 May 1]. In: StatPearls [Internet]. Treasure Island (FL): StatPearls Publishing; 2024 Jan-. Retrieved from: <https://www.ncbi.nlm.nih.gov/books/NBK557536/>
38. Lohr GW, Waller HD. (1974). Glucose-6-Phosphate Dehydrogenase. Methods of enzymatic analysis. 3<sup>rd</sup> Edition – Verlag Chemie, Wehnhelm; 636.
39. Ellman GL. (1959). Tissue sulphhydryl groups. *Arch Biochem Biophys*, 82:70-77.
40. Rotruck JT, Pope AL, Ganther HE, Swanson AB, Hafeman DG, Hoekstra WG. (1973). Selenium: biochemical role as a component of glutathione peroxidase. *J Sci*. 179 (4073): 588–590.
41. Ogbole FA, Ayakoroma JJ. (2024). Effect of Diet on Interleukin-2 Gene Expression and Hemacrit. *International Journal of Research Publication and Reviews*, 5(8):1417-1426.
42. Ogbole FA, Igwe CU, Onuoha HC, Nzebude CP. (2023). Characterization of ABO / Rhesus Antigen Polymorphism Associated with Malaria in a Malaria Hotspot in Bayelsa State, Niger Delta, Nigeria. *IJBRR*, 32(7): 42-52.
43. Parveen S, Bharose R, Singh D. (2017). Effect of tannery waste water on lactate dehydrogenase (LDH) enzyme activity of fresh water fish, *Channa punctatus*. *JEZS*, 5(2): 643-647.
44. Panepucci L, Fernandes MN, Sanches JR, Rantin FT. (2000). Changes in lactate dehydrogenase and malate dehydrogenase activities during hypoxia and after temperature acclimation in the armored fish, *Rhinelepis strigosa* (Siluriformes, Loricariidae). *Rev. Brasil. Biol*, 60(2): 353-360.

45. Naglaa E, Bahnasawy M. (2019). Comparative and interactive biochemical effects of sub-lethal concentrations of cadmium and lead on some tissues of the African Catfish (*Clarias gariepinus*). *Toxicol. Res*, 35(3):249-255.
46. Rusinova OS. (2000). Activity of glucose-6-phosphate dehydrogenase in black sea fish erythrocytes. *J EvolBiochem Phys*, 36:138–142.
47. Saif M, Al-Ghais (2013), Acetylcholinesterase, glutathione and hepatosomatic index as potential biomarkers of sewage pollution and depuration in fish. *Mar Pollut Bull*, 74(1):183-186.
48. Li X, Lin L, Luan T, Yang L, Lan C. (2008). Effects of landfill leachate effluent and bisphenol A on glutathione and glutathione-related enzymes in the gills and digestive glands of the freshwater snail *Bellamyapurificata*, *Chemosphere*, 70(10):1903-1909.
49. Javed M, Ahmad I, Usmani N, Ahmad M. (2016). Bioaccumulation, oxidative stress and genotoxicity in fish (*Channa punctatus*) exposed to a thermal power plant effluent. *Ecotoxicol Environ Saf*. 127:163-169.
50. Cazenavea J, Bacchettaa C, Rossi A, Ale A, Campanaa M, Parmaa MJ. (2014). Deleterious effects of wastewater on the health status of fish: A field caging study. *Ecol Indic*, 38:104–112
51. Mustafa SA, Al-RudainyAJ, Salman NM. (2024). Effect of environmental pollutants on fish health: An overview. *J. Aquat Res*, 50(2): 225-233.