

## Review Article

# PHYTOREMEDIATION: AN EFFECTIVE WAY TO TREAT HEAVY METAL CONTAMINATION- A REVIEW

### Abstract

Over the years there has been a drastic increase in heavy metal contamination in the soil due to a number of natural and industrial processes. As these heavy metals are non-biodegradable in nature, they remain in the environment for long periods of time and may infiltrate into the food chain through plants and eventually get accumulated in the human body *via* biomagnification. Because of their poisonous nature, they pose a major threat to human health and environment. Therefore, the process of cleaning up of this contamination from the environment is of great significance. Phytoremediation, an ecologically viable process, can be a very good mitigation strategy to rid the soil from this heavy metal pollution. Though phytoremediation is not capable of completely removing dangerous pollutants, it has no deleterious effect on the ecosystem as it is an environment friendly, cost-effective, and natural process. Different plants, the majority of which belong to the *Brassicaceae* family, that are used in the phytoremediation process are referred to as hyper accumulators. The efficacy of phytoremediation can be improved by having deeper knowledge and understanding of different mechanisms contributing to heavy metal accumulation and tolerance in plants. Different mechanisms such as phytoextraction, phytostabilization, phytovolatilization, and rhizofiltration have been employed to reduce contamination of heavy metals in the soil. Among these, the first and second mechanisms are the most dependable. A number of factors are taken into consideration for choosing the most appropriate phytoremediation strategy for soil decontamination. There are a number of factors that influence the phytoremediation process which include the types of contaminants, their characteristics and the selection of plant species. Other factors to consider include climatic conditions, flooding and ageing, the effect of salt on the process, soil variables and the redox potential.

**Keywords:** Bioavailability, Heavy metals, Hyperaccumulators, Phytoremediation, Permissible limits, Toxicity

## **Introduction**

As a result of rapid urbanization and industrialization in recent years, the concentration of heavy metals in the environment has increased significantly. This is a global concern that warrants attention and action (Ashraf *et al.*, 2019). These heavy metals include cadmium (Cd), arsenic (As), mercury (Hg), lead (Pb), zinc (Zn), copper (Cu), chromium (Cr) and nickel (Ni). The accumulation of these heavy metals/metalloids in the environment results from a variety of natural and anthropogenic sources like produced water from the oil and gas industry (Pichtel, 2016), phosphate fertiliser in agriculture (Hamzahet *al.*, 2016), sewage sludge (Farahat and Linderholm, 2015), metal mining and smelting (Muradogluet *al.*, 2015). Metal contamination in the environment is a worldwide problem, they are non-degradable and remain in the soil for a considerably long period of time thereby posing a great threat to the environment (Sumanet *al.*, 2018). They find their entry into the food chain through crops and assimilate in the human body through biomagnification, posing a serious health risk.

It was Andrea Cesalpino in the 16th century, who discovered phytoremediation (Brooks, 1998). Phytoremediation is a process of naturally removing harmful metals from the environment via plants. It is environment friendly with low operating cost when compared to alternative manual procedures like electrokinetic soil remediation and acid leaching or natural methods (ion exchange, membrane filtration and adsorption) (Renuet *al.*, 2017). Even at low concentration, plants, through their root system can absorb heavy metal ionic compounds from the soil. By extending their root systems into the soil matrix, plants form a rhizospheric ecosystem, accumulating heavy metals and modulating their bioavailability, restoring soil fertility and thereby reclaiming polluted soil (DalCorsoet *al.*, 2019). Plants that can accumulate metals to a high concentration are referred as "hyperaccumulators" (Visoottiviset *et al.* 2002).

## **Source, Effect and Limit of Different Harmful Metals**

Heavy metals are brought into the soil environment by both geogenic and anthropogenic processes. Heavy metals from parent soil materials are released into the soil ecosystem by natural weathering processes; however they are not readily available for uptake by plants. The contamination of ground water with arsenic at a large scale in Bangladesh and West Bengal (India) is a very good example of geogenic contamination (Mahimairajaet *al.*, 2005). The

contribution of anthropogenic source to heavy metal pollution has risen drastically in the recent years as a result of giant strides in industrialization, urbanization, lavish lifestyles and population growth (Asati *et al.*, 2016). Agricultural activities in areas on the outskirts of industrial units and peri-urban localities are being critically scrutinized due to high risk of metal pollution from sewage effluents and sludge produced from these units, which have the potential to enter the food chain through crops grown in contaminated fields, particularly vegetables. (Purakayastha, 2007, Mitra and Gupta, 1999). Chemical fertilizers also introduce into the soil a significant amount of heavy metals, for example, phosphate compounds, contain metals such as Cd. In many formulations of fungicides, such as Copper oxychloride and Bordeaux mixture, Cu is used as a trace element nutrient. It is also used as a growth promoter in poultry and piggery units in horticulture, agriculture and animal industries.

Bioavailability of these metals, on the other hand, is influenced by a variety of factors which include soil quality, exposure pathways, and animal physiological traits, and can vary from one organism to the next (Luoma *et al.*, 2005). These toxic metals can alter the water and nutrient absorption balance, interfere with their transport to aboveground plant parts and cause adverse effects on shoot growth, thereby stifle the overall plant growth (Popova *et al.*, 2009). Some metals like Cu and Zn, act as activators and cofactors of an enzyme's normal function (Mildvan, 1970). On the other hand, toxic metals like Cd, Hg, As and lead are harmful to plants and all living things.

### **Toxicity of Heavy Metals**

At high concentrations, all the heavy metals are toxic to crop plants. The deficiency of essential heavy metal elements is more common in agriculture than their toxicity. Heavy metals interfere with physiological and biochemical processes, reducing growth, degrading cell organelles, and preventing photosynthesis thereby, poisoning macro- and microorganisms. Direct ingestion of contaminated soils, consumption of crops and vegetables grown on contaminated lands, or drinking water that has percolated through contaminated soils can expose humans and ecosystems to heavy metals (lead, chromium, arsenic, zinc, cadmium, copper, mercury, and nickel) (McLaughlin *et al.*, 2000). For example, Chaney *et al.* (2005) concluded that lives of subsistence farmers who eat rice grown on contaminated sites are at risk of cadmium poisoning. According to Kuzovkina *et al.* (2004), cadmium is highly phytotoxic and can result in rapid death. According to a report released by a US environmental action group, the world's most

polluted places endanger the health of more than 10 million people in many countries (Chhotu *et al.* 2009). On the basis of their function in biological systems, heavy metals are classified into two categories *viz.*, essential and non-essential. The essential group comprises of Cu, Fe, Mn, Ni, and Zn. These are important for physiological and biochemical processes during the plant life cycle; however, they can be toxic, when present in excess concentrations. Pb, Cd, As, and Hg fall into non-essential category having no known function in plants and are highly toxic (Fasani *et al.*, 2018). They interfere with metabolic processes and inhibit growth at higher concentrations, which can lead to plant death. They disrupt a number of physiological and biochemical processes in crop plants, resulting in reduction of agricultural productivity (Clemens, 2006). Phytotoxicity due to heavy metals can result from changes in cellular and molecular physiological processes due to deactivation of enzymes or the functional group blocking of metabolically important molecules (Ahmadpouret *et al.*, 2012). Some heavy metals like Pb tend to accumulate more in the roots than in stems, because some barriers prevent their movement from the roots to aboveground parts. However, metals like Cd, move around in plants more easily (Garbisu and Alkorta, 2001). Cadmium damages the plant's light harvesting complex II, as well as photosystems I and, which are involved in photosynthesis. Cadmium treatment reduces total chlorophyll content in the plants simultaneously increasing non-photochemical quenching in *Brassica napus*. It is likely to cause obstruction in the movement of  $K^+$ ,  $Ca^{2+}$  and abscisic acid in guard cells, as well as stomatal opening (Shaw, 1995).

### **Hyperaccumulators**

In order to survive in environments with high metal content, naturally growing plants use two primary tactics as part of their survival strategy. The majority of the plant species which are capable of growing in soils containing increased concentrations of hazardous trace elements have adopted the survival strategy of maximal exclusion of heavy metal (HM) ions from the plant. They are known as Excluders. Whenever an HM ion is taken up by these plants, the poisonous effect is limited to the roots, where it is detoxified, whereas the aboveground parts remain more or less unharmed. On the other hand, some plants, when exposed to higher concentration of HM, are able to collect them in their aerial parts without phytotoxic symptoms and are called hyperaccumulators (Van der Ent *et al.*, 2013). Initially the term hyperaccumulator was applied when describing the New Caledonian Ni accumulating tree *Sebartia acuminata*, the concentration of nickel in the latex on dry weight basis was determined to be 26 percent (Jaffré *et al.*, 1976).

For plants to be categorized into hyperaccumulators, the concentration limitations of several metal elements in their dry biomass are 10,000 mgkg<sup>-1</sup> for Zn and Mn, 1,000 mgkg<sup>-1</sup> for Co, Cu, Ni and Pb, and 100 mgkg<sup>-1</sup> for Cd and Se (Baker and Brooks, 1989). These values are up to 100–1,000 times more than for non-hyperaccumulating species under the same conditions (Rascio & Navari-Izzo, 2011). The number of recognized hyper-accumulators has been steadily rising, with over 450 HM-hyperaccumulating species known as of 2015, found in 45 angiosperm families. About 25 percent of hyperaccumulators identified so far recruit from the family *Brassicaceae*; other families rich in hyperaccumulators include *Asteraceae*, *Euphorbiaceae*, *Rubiaceae*, *Fabaceae*, *Scrophulariaceae*, *Myrtaceae*, *Proteaceae*, *Caryophyllaceae*, *Tiliaceae*, etc. (Rascio and Navari-Izzo, 2011). According to Brooks *et al.*, (1979), some plant genera appear to feature a large proportion of hyperaccumulators e.g., while assessing the genus *Alyssum*, out of 170 species tested, 48 hyperaccumulated Ni. The ability to hyperaccumulate has been discovered to vary within a species; metal specificity and accumulation can vary among different populations (Assuncao *et al.*, 2008). According to Milner and Kochian (2008), who evaluated the Zn, Cd and Ni model hyperaccumulator alpine pennycress (*Nocca caerulea*), several essential physiological stages of heavy metal detoxification in hyperaccumulators are different as compared to non-hyperaccumulators. These are:

1. Increased uptake of HM ions from the rhizosphere across the root cell plasma membrane.
2. Reduced sequestration of HM ions in root vacuoles.
3. Intensified loading of HM ions into the xylem for transportation to shoots.
4. Stimulated influx of HM ions across the plasma membrane of leaf cell.
5. Sequestration in the leaf vacuole.

Nevertheless, in the sequestration of HM ions within a plant body, a major involvement of cell wall components, especially low-methylesterified pectins, was also hypothesized by Krzesłowska, 2011. The basic mechanism for hyperaccumulation in some plants is due to improved active metal transport rather than strengthened metal complexation by some intracellular ligands like phytochelatin, glutathione, or metallothionein (Leitenmaier and Küpper, 2013). Further, in case of extremophilic Zn- and Cd-hyperaccumulator *Arabidopsis halleri*, transcriptomic and genome-wide analysis showed an enrichment in HM-transporting P-ATPase HMA4 gene copies and corresponding transcripts as well as other transition metal

homeostasis and biotic stress function genes, as compared to the non-accumulating sister species *Arabidopsis lyrata* and closely related reference model species *Arabidopsis thaliana* (Suryawanshi *et al.*, 2016). More detailed information on the various tolerance strategies of plants under HM stress and on hyperaccumulators, their taxonomic and geographical distribution, metal specificity as well as molecular mechanisms behind hyperaccumulation have been provided by Rascio and Navari-Izzo (2011); Leitenmaier and Küpper (2013) and van der Ent *et al.* (2013).

## **Techniques of Phytoremediation**

### **Phytoextraction**

A type of phytoremediation in which the plant extract harmful elements or compounds from the soil or water is called Phyto-extraction. It is basically the ability of some plants to absorb metals from the soil and assimilate them in the harvestable shoots. The concentration of heavy metals is higher in roots than the shoots, but there is a limit to this concentration. For example, root uptake of Pb increased in hydroponically grown plants with increase in its concentration in the solution and reached maximum, thereafter did not increase any further with increase in the concentration (Deepa, *et al.* 2006). The percentage of exchangeable Cd in the soil decreased as maize phytoextraction increased (Zhang *et al.*, 2009). Phytoextraction could become a commercial technology in the future.

### **Phytodegradation**

Phytodegradation, also known as Phytotransformation, involves the use of plants and microorganisms for uptake, metabolizing and degrading the organic contaminants. Plant roots in conjunction with microorganisms are used to detoxify organic contaminants in the soil (Garbisu and Alkorta, 2001). Various enzymes produced by certain plants help to metabolise contaminants that may be released into the rhizosphere, where they can persist (Singh *et al.*, 2003). Some halophytes convert Cr (VI) to less toxic Cr (III) (Cacador and Duarte, 2015). Various bacterial and fungal microorganisms can also help in conversion of toxic metals into less toxic forms. By producing certain specific enzymes, some plants can decontaminate soil, sludge, sediment, and ground and surface water (Pivetz, 2001).

### **Phytostabilization**

It is the process of reducing heavy metal mobility in the soil. The immobilization can be achieved by decreasing contaminant solubility or bioavailability in the food chain, minimizing soil erosion and reducing windblown dust. Revegetation of the site significantly improves

physicochemical and biological properties of the contaminated soil by increasing organic matter content, cation exchange capacity, nutrient levels and biological activity (Arie, *et al.* 2004). Heavy metals like Cd, Cu, Zn, Cr and As can all be remedied using phytostabilization. The most important advantage of this technology is that it does not involve the disposal of hazardous materials (United States Protection Agency, 2000) and it is also very effective when quick immobilization of ground and surface waters is needed (Chhotu, *et al.* 2009).

### **Rhizofiltration**

The technique to remediate contaminated groundwater, surface water, and wastewater with low levels of contaminants through the use of various plants is called Rhizofiltration. Ultramatic soils enriched with Cd, Ni, Zn, or Pb can be remedied by growing *Berkheyacoddii* plants, which accumulate a significant amount of these metals. Exposed shoots as compared excised shoots in solutions containing the same heavy metals accumulate a higher amount of these metals in the leaves (Mesjasz-Przybylowicz *et al.*, 2004). The ability of various plants like sunflower, rye, spinach, Indian mustard, tobacco and corn to remove lead from water has been studied, with sunflower having the greatest ability (Chhotu *et al.* 2009). Heavy metals like Pb, Cd, Cr, Zn, Cu and Ni which are primarily retained within the roots can be effectively treated with rhizofiltration (United States Protection Agency, 2000).

### **Phytovolatilization**

This involves the usual uptake and transpiration of a contaminant by a plant, followed by the release of the contaminant or its modified form into the atmosphere. A number of these contaminants can pass through the plants to their leaves and volatilize at a low concentration into the atmosphere (Ismail, 2012). Some plants grown in high selenium media are able to transform Se into dimethylselenide and dimethyldiselenide (Banuelos, 2000). Some bacteria are capable to absorb mercury and evaporate it. In contrast to other remediation techniques, contaminants once removed by volatilization, cannot be prevented from spreading to other areas.

### **Factors that Influence Heavy Metal Phytoavailability**

The terms phytoavailability and bioavailability are used to characterise the extent to which pollutants are available for uptake or absorption by living organisms. Only the part which is "phytoavailable" to them causes plants to respond (Chang *et al.* 2014). Bioavailability of metals

in soils is a critical element governing heavy metal uptake by plants in phytoremediation especially phytoextraction (Faridet *al.* 2013). On the other hand, metal phytoavailability, is a complex phenomenon that is affected by a number of interconnected elements.

## **Soil Characteristics**

### **Soil pH**

Soil pH has a direct impact on metal phytoavailability, as soil acidity determines metal solubility and their capacity to move in the soil solution. Metal cations are the most mobile under acidic soil conditions, whereas anions have the tendency to sorb to oxide minerals in this pH (Dzombak and morel, 1987). Metal bioavailability increases at low pH as more metals are liberated into the soil solution, due to competition with  $H^+$  ions. At high pH, cations adsorb to mineral surfaces or precipitate, causing immobilization of metal anions. Majority of the heavy metals in soil are not available to plants at neutral or alkaline pH, especially Cr and Pb, which are intrinsically immobile (Mahmood, 2010).

### **Soil Texture**

The texture of the soil provides information regarding the concentration of particles like oxides and clay (Sherene, 2010). The particle size distribution actively affects the level of metal contamination. The reactivity and surface area of fine particles is higher than the coarser ones, resulting in greater pollution levels in the fine fraction of soil (Evanko and Dzombak, 2004). For instance, Pb levels in fine and coarse textured soils have been recorded to be  $3889 \text{ mg kg}^{-1}$  and  $530 \text{ mg kg}^{-1}$ , respectively (Sherene, 2010).

### **Soil Organic Matter**

The behavior of heavy metals in the soil is being regulated at a large scale by the presence of organic matter in the soil. Organic matter is that fraction of the soil that may diminish the ability metals to be phytotoxic due to metal-organic complexation. Cation exchange capacity of the soil is enhanced by organic carbon, thereby allowing more nutrient retention by the plants (Yobouetet *al.*, 2010). Therefore, trace metal uptake by the plants can be reduced to a great degree by increasing the organic matter content in the soil. The soils rich in organic matter actively maintain metallic elements. The bioavailabilities of metals can be reduced in contaminated soils containing labile elements by compost amendments through sorption processes (Brown *et al.*, 2003).

### **Redox Potential**

One of the most important soil attributes that influences metal speciation variations is the redox potential. It is determined by the microbial activity that causes oxidation-reduction reactions in the soil. These redox processes convert contaminants into such forms that are non-hazardous or less toxic, more stable, less mobile, and/or inactive (Alkorta *et al.*, 2004).

## **Conclusion**

The process of phytoremediation seems to be an efficient and promising cleanup strategy for a wide variety of contaminants (organic and inorganic) and sites. It can be used both in-situ and ex-situ, depending on the situation. In-situ application is usually preferred as it limits the soil and ecosystem disturbance and pollution spread by air and waterborne wastes. When executed properly, phytoremediation has the advantage of being ecologically beneficial and visually acceptable to the general public. It can be executed in a simple manner not requiring expensive equipment or highly qualified personnel. The main advantage of phytoremediation is that it is cost effective compared to traditional clean-up solutions. For instance, the cost of cleaning up one acre of sandy loam soil with a contamination depth of 50 cm through plants was estimated to be \$60,000-\$100,000 compared to \$400,000 in case of traditional excavation and disposal approach (Number Ma *et al.* 1997). Therefore, the current review focuses on different phytoremediation processes and their potential as remediation techniques that take advantage of plant's inherent ability to remove contaminants from the environment, but which are still not commercially available in many regions of the world.

## **References**

- Ahmadpour, P., Ahmadpour, F., Mahmud, T.M.M., Arifin Abdul, H., Soleimani, M., Hosseini Tayefeh, F. (2012) Phytoremediation of heavy metals: A green technology. *African Journal of Biotechnology*, 11(76), 14036- 14043
- Alkorta, I., Hernandez-Allica, J., Becerril, J.M., Amezaga, I., Albizu, I. and Garbisu, C. (2004) Recent Findings on the Phytoremediation of Soils Contaminated with Environmentally Toxic Heavy Metals and Metalloids Such as Zinc, Cadmium, Lead, and Arsenic. *Reviews in Environmental Science and Bio/Technology*, 3, 71-90.
- Asati, A.; Phichhole, M.; Nikhil, K. Effect of Heavy Metals on Plants: An Overview *International Journal of Application or Innovation in Engineering & Management (IJAIEM)*. Int. J. Appl. Innov. Eng. Manag. 2016, 5, 2319–4847.

Ashraf, S., Ali, Q., Zahir, Z. A., Ashraf, S., and Asghar, H. N. (2019). Phytoremediation: environmentally sustainable way for reclamation of heavy metal polluted soils. *Ecotox. Environ. Safe.* 174, 714–727.

Assuncao, A. G. L., Bleeker, P., Ten Bookum, W. M., Vooijs, R., and Schat, H. (2008). Intraspecific variation of metal preference patterns for hyperaccumulation in *Thlaspi caerulescens*: evidence from binary metal exposures. *Plant Soil* 303, 289–299.

Baker, A. J. M., and Brooks, R. R. (1989). Terrestrial higher plants which hyperaccumulate metallic elements—a review of their distribution, ecology and phytochemistry. *Biorecovery* 1, 81–126.

Banuelos, G.S. Phytoextraction of selenium from soils irrigated with selenium-laden effluent. *Plant Soil.* 2000; 224(2):251-258.

Brooks, R. R., Morrison, R. S., Reeves, R. D., Dudley, T. R., and Akman, Y. (1979). Hyperaccumulation of nickel by *Alyssum linnaeus* (Cruciferae). *Proc. R. Soc. B Biol. Sci.* 203, 387–403.

Brooks, R.R. Phytoremediation by volatilisation. In *Plants that Hyper-Accumulate Heavy Metals: Their Role in Phytoremediation, Microbiology, Archaeology, Mineral Exploration and Phytomining*; CAB International: Wallingford, UK, 1998; pp. 289–312.

Brown, S.L., Henry, C.L., Chaney, R., Compton, H. and De Volder, P.S. (2003) Using Municipal Biosolids in Combination with Other Residuals to Restore Metal-Contaminated Mining Areas. *Plant Soil*, 249, 203-215

Caçador I, Duarte B. Chromium Phyto-transformation in Salt Marshes: The Role of Halophytes. *Phytoremediation*, 2015, 211-217.

Chaney RL, Reeves PG, Ryan JA, Simmons RW, Welch RM, Angle JS. An improved understanding of soil Cd risk to humans and low cost methods to phytoextract Cd from contaminated soils to prevent soil Cd risks. *Bio Metals.*2005; 17:549-553.

Chang, Y.T., Hseu, Z.Y. and Zehetner, F. (2014) Evaluation of Phytoavailability of Heavy Metals to Chinese Cabbage (*Brassica chinensis* L.) in Rural Soils. Scientific World Journal, 2014, Article ID: 309396.

Chhotu D, Jadia D, Fulekar MH. Phytoremediation of heavy metals: Recent techniques. African Journal of Biotechnology. 2009; 8(6):921-92.

Clemens, S. (2006). Toxic metal accumulation, responses to exposure and mechanisms of tolerance in plants. *Biochimie* 88, 1707–1719. doi: 10.1016/j. biochi.2006.07.003

DalCorso, G., Fasani, E., Manara, A., Visioli, G., and Furini, A. (2019). Heavy metal pollutions: state of the art and innovation in phytoremediation. *Int. J. Mol. Sci.* 20:3412. doi: 10.3390/ijms20143412

Deepa KK, Sathiskumar M, Binupriya AR, Murugesan GS, Swaminathan K, Yun SE. Sorption of Cr (VI) from dilute solutions and wastewater by live and pretreated biomass of *Aspergillusflavus*. *Chemosphere*.2006; 62:833-840.

Dzombak, D.A. and Morel, F.M.M. (1987) Adsorption of Inorganic Pollutants in Aquatic Systems. *Journal of Hydraulic Engineering*, 113, 430-475

Farahat, E., and Linderholm, H. W. (2015).The effect of long-term wastewater irrigation on accumulation and transfer of heavy metals in *Cupressus sempervirens* leaves and adjacent soils. *Sci. Total Environ.* 51, 1–7.

Farid, M., Ali, S., Shakoor, M.B., Bharwana, S.A., Rizvi, H., Ehsan, S., Tauqeer, H.M., Iftikhar, U. and Hannan, F. (2013) EDTA Assisted Phytoremediation of Cadmium, Lead and Zinc. *International Journal of Agronomy and Plant Production*, 4, 2833-2846.

Fasani, E., Manara, A., Martini, F., Furini, A., and DalCorso, G. (2018). The potential of genetic engineering of plants for the remediation of soils contaminated with heavy metals. *Plant Cell Environ.* 41, 1201–1232. doi: 10. 1111/pce.12963

Garbisu C, Alkorta I. Phytoextraction: A cost-effective plant-based technology for the removal of metals from the environment. *Bioresour. Technol.* 2001; 77(3):229- 236.

Hamzah, A., Hapsari, R. I., and Wisnubroto, E. I. (2016). Phytoremediation of Cadmium-contaminated agricultural land using indigenous plants. *Int. J. Environ. Agric. Res.* 2, 8–14.

Jaffré, T., Brooks, R. R., Lee, J., and Reeves, R. D. (1976). *Sebertia acuminata*: a hyperaccumulator of nickel from New Caledonia. *Science* 193, 579–580.

Krzyszowska, M. (2011). The cell wall in plant cell response to trace metals: polysaccharide remodeling and its role in defense strategy. *Acta Physiol. Plant.* 33, 35–51.

Kuzovkina YA, Knee M, Quigley MF. Cadmium and copper uptake and translocation in five Willow *salix* L. species. *Int. J. Phytoremediation.* 2004; 6:269-287.

Leitenmaier, B., and Küpper, H. (2013). Compartmentation and complexation of metals in hyperaccumulator plants. *Front. Plant Sci.* 4:374.

Luoma, S.; Rainbow, P.S. Why is metal bioaccumulation so variable? Biodynamics as a unifying concept. *Environ. Sci. Technol.* 2005, 39, 1921–1931.

Mahimairaja, S., Bolan, N.S., Adriano, D.C., Robinson, B., 2005. Arsenic contamination and its risk management in complex environmental settings. *Advances in Agronomy* 86, 1-82

Mahmood, T. (2010) Phytoextraction of Heavy Metals—The Process and Scope for Remediation of Contaminated Soils. *Soil & Environment*, 29, 91-109.

McLaughlin MJ, Zarcinas BA, Stevens DP, Cook N. Soil testing for heavy metals. *Commun. Soil. Sci. Plant Anal.* 2000; 31(11-14):1661-1700.

Mesjasz-Przybyłowicz J, Nakonieczny M, Migula P, Augustyniak M, Tarnawska M, Reimold WU et al. Uptake of cadmium, lead, nickel and zinc from soil and water solutions by the nickel hyperaccumulator *Berkheyacodii*. *Acta Biol. Cracov. Bot.* 2004; 46:75-85.

Mildvan, A.S. 9 Metals in Enzyme Catalysis. *Enzymes* 1970, 2, 445–536.

Milner, M. J., and Kochian, L. V. (2008). Investigating heavy-metal hyperaccumulation using *Thlaspi caerulescens* as a model system. *Ann. Bot.* 102, 3–13.

Mitra, A., Gupta, S.K., 1999. Effect of sewage water irrigation on essential plant nutrient and pollutant element status in a vegetable growing area around Calcutta. *Journal of the Indian Society of Soil Science* 47, 99-105.

Muradoglu, F., Gundogdu, M., Ercisli, S., Encu, T., Balta, F., Jaafar, H. Z. (2015). Cadmium toxicity affects chlorophyll a and b content, antioxidant enzyme activities and mineral nutrient accumulation in strawberry. *Biol. Res.* 48:11.

Number Ma, W. A., C., J. Kingscott and M. Evans (1997). "Recent Developments for In Situ Treatment of Metal Contaminated Soils."

Pichtel, J. (2016). Oil and gas production wastewater: soil contamination and pollution prevention. *Appl. Environ. Soil Sci.* 2016:2707989.

Pivetz BE. Ground Water Issue: Phytoremediation of contaminated soil and ground water at hazardous waste sites, 2001, 1-36.

Popova, L.P.; Maslenkova, L.T.; Yordanova, R.Y.; Ivanova, A.P.; Krantev, A.P.; Szalai, G.; Janda, T. Exogenous treatment with salicylic acid attenuates cadmium toxicity in pea seedlings. *Plant Physiol. Biochem.* 2009, 47, 224–231.

Purakayastha, T.J., 2007. Phytoremediation-the green cure technology for amelioration of heavy metal contaminated soils. *Indian Farming* 17-19 & 26-30.

Rascio, N., and Navari-Izzo, F. (2011). Heavy metal hyperaccumulating plants: how and why do they do it? and what makes them so interesting? *Plant Sci.* 180, 169–181. doi: 10.1016/j.plantsci.2010.08.016

Renu, N.A.; Agarwal, M.; Singh, K. Methodologies for removal of heavy metal ions from wastewater: An overview. *Interdiscip. Environ. Rev.* 2017, 18, 124.

Shaw BP. Effects of mercury and cadmium on the activities of antioxidative enzymes in the seedlings of *Phaseolusaureus*. *BiologicaPlantarum.* 1995; 37:587- 596.

Sherene, T. (2010) Mobility and Transport of Heavy Metals in Polluted Soil Environment. *Biological Forum—An International Journal*, 2, 112-121.

Singh OV, Labana S, Pandey G, Budhiraja R, Jain RK. Phytoremediation: an overview of metallic ion decontamination from soil. *Appl. Microbiol. Biotechnol.* 2003; 61(5):405-412.

Suryawanshi, V., Talke, I. N., Weber, M., Eils, R., Brors, B., Clemens, S. (2016). Between-species differences in gene copy number are enriched among functions critical for adaptive evolution in *Arabidopsis halleri*. *BMC Genomics* 17: 1034. doi: 10.1186/s12864-016-3319-5

USEPA (United States Environmental Protection Agency). Electrokinetic and Phytoremediation In Situ Treatment of Metal-Contaminated Soil: State of the Practice. Draft for Final Review. EPA/542/R-00/XXX. US Environmental Protection Agency, Office of Solid Waste and Emergency Response Technology Innovation Office, Washington, DC, 2000.

Visoottiviseth P, Francesconi K, Sridokchan W. The potential of Thai indigenous plant species for the phytoremediation of arsenic contaminated land. *Environmental Pollution*. 2002; 118:453-461.

Yobouet, Y.A., Adouby, K., Trokourey, A. and Yao, B. (2010) Cadmium, Copper, Lead and Zinc Speciation in Contaminated Soils. *International Journal of Engineering Science and Technology*, 2, 802-812.

Zhang H, Dang Z, Zheng LC, Yi XY. Remediation of soil co-contaminated with pyrene and cadmium by growing maize (*Zea mays* L.). *Int. J Environ. Sci. Tech.* 2009; 6:249-258